

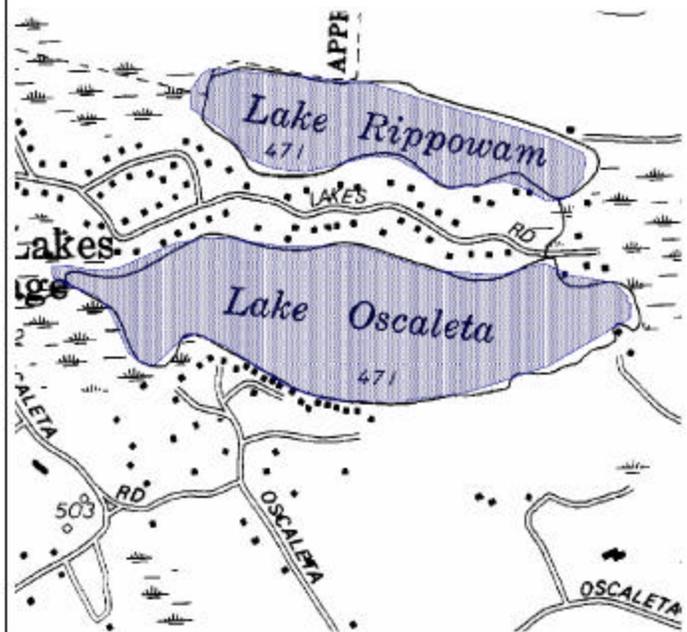
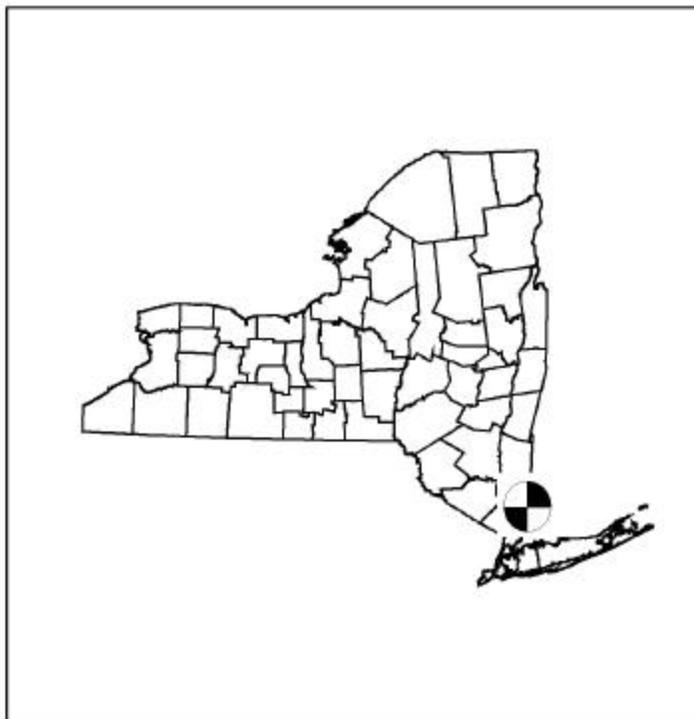


Division of Water

**New York
Citizens Statewide Lake Assessment Program
(CSLAP)**

2006 Annual Report- Lake Oscaleta

September, 2007



2006 INTERPRETIVE SUMMARY

NEW YORK CITIZENS STATEWIDE LAKE ASSESSMENT PROGRAM (CSLAP)

LAKE OSCALETA

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NY Federation of Lake Associations

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BACKGROUND AND ACKNOWLEDGMENT

The Citizens Statewide Lake Assessment Program (CSLAP) is a volunteer lake monitoring program conducted by the NYS Department of Environmental Conservation (NYSDEC) and the NYS Federation of Lake Associations (FOLA). Founded in 1986 with 25 pilot lakes, the program has involved more than 200 lakes, ponds, and reservoirs and 1000 volunteers from eastern Long Island to the Northern Adirondacks to the western-most lake in New York, and from 10 acre ponds to several Finger Lakes, Lake Ontario, Lake George, and lakes within state parks. In this program, lay volunteers trained by the NYSDEC and FOLA collect water samples, observations, and perception data every other week in a fifteen-week interval between May and October. Water samples are analyzed by certified laboratories. Analytical results are interpreted by the NYSDEC and FOLA, and utilized for a variety of purposes by the State of New York, local governments, researchers, and, most importantly, participating lake associations. This report summarizes the 2006 sampling results for **Lake Oscaleta**.

Lake Oscaleta is a 65 acre, class B lake found in the Town of Lewisboro in Westchester County, just north of the New York City region of New York State. Lake Oscaleta was first sampled as part of CSLAP in 2006. The following volunteers have participated in CSLAP, and deserve most of the credit for the success of this program at **Lake Oscaleta: Paul Lewis, Shannon Robinette, Dick Karl, Barbara Posner, Paul Lewis, and Leslie Daley.**

In addition, the authors wish to acknowledge the following individuals, without whom this project and report would never have been completed:

From the Department of Environmental Conservation, N.G. Kaul, Sal Pagano, Dan Barolo, Italo Carcich, Phil DeGaetano, Dick Draper, and Jeff Myers for supporting CSLAP for the past twenty years; Jay Bloomfield and James Sutherland, for their work in developing and implementing the program; and the technical staff from the Lake Services Section, for continued technical review of program design.

From the Federation of Lake Associations, Anne Saltman, Dr. John Colgan, Don Keppel, Bob Rosati, Don Cook, Nancy Mueller and the Board of Directors, for their continued strong support of CSLAP.

The New York State Department of Health (prior to 2002), particularly Jean White, and Upstate Freshwater Institute (since 2002), particularly Steve Effler and Jennifer Aicher, provided laboratory materials and all analytical services, reviewed the raw data, and implemented the quality assurance/quality control program.

Finally, but most importantly, the authors would like to thank the more than 1000 volunteers who have made CSLAP a model for lay monitoring programs throughout the country and the recipient of a national environmental achievement award. Their time and effort have served to greatly expand the efforts of the state and the public to protect and enhance the magnificent water resources of New York State.

LAKE OSCALETA FINDINGS AND EXECUTIVE SUMMARY

Lake Oscaleta was sampled as part of the New York Citizens Statewide Lake Assessment Program in 2006. For all program waters, water quality conditions and public perception of the lake each year and historically have been evaluated within annual reports issued after each sampling season. This report attempts to summarize both the 2006 CSLAP data and an historical comparison of the data collected within the 2006 sampling season and data collected at Lake Oscaleta prior to 2006.

The majority of the short- and long-term analyses of the water quality conditions in Lake Oscaleta are summarized in Table 2, divided into assessments of eutrophication indicators, other water quality indicators, and lake perception indicators. Although eight samples were collected in 2006, water quality assessments must be considered preliminary, since these data may be variable from year to year. The CSLAP data indicate that the lake is best classified as eutrophic, or highly productive. This assessment is consistent whether evaluating nutrient, or algae levels, although water clarity readings may be more typical of moderately productive (mesotrophic) lakes. The nitrogen to phosphorus ratios indicate that algae levels in Lake Oscaleta are probably controlled by phosphorus, and it is likely that phosphorus inputs need to be addressed to improve water clarity and prevent algal blooms. Lake productivity does not vary in any predictable way over the summer. Phosphorus levels in the lake regularly exceed the state phosphorus guidance value, although water transparency readings only rarely fail to reach the minimum recommended water clarity for swimming beaches. Deepwater phosphorus readings are higher than those measured at the lake surface, suggesting that internal nutrient cycling (release of phosphorus from bottom sediments to the deep waters, and then eventually into the surface waters) may be significant. In short, Lake Oscaleta appears to be a highly productive lake, and thus may regularly suffer from algal blooms and or reduced water clarity.

The lake is moderately colored (intermediate levels of dissolved organic matter) and it is likely that these readings reflect the soil and vegetation characteristics of the watershed (i.e. “natural” conditions at the lake). Color readings are probably not high enough to influence the water transparency, and water clarity is more likely to be affected by high algae levels. The lake has moderately soft water, alkaline (above neutral) pH readings, and low nitrate and ammonia readings. Neither nitrate nor ammonia levels appear to warrant a threat to the lake. pH readings occasionally exceed the NYS water quality standards (=6.5 to 8.5), but are probably adequate to support most aquatic organisms. Conductivity readings varied during the summer, but all readings are typical of moderately softwater lakes. Calcium levels are high enough to support zebra mussel growth, but it is not believed that zebra mussels have been found in the lake. Additional data will help to determine if these data are typical of conditions in Lake Oscaleta.

The recreational suitability of Lake Oscaleta was most often “slightly” impaired. The lake was most often described as (having) “definite algal greenness”, with mostly sub-surface growth of aquatic plants. This recreational assessment is about as favorable as in other lakes with similar water quality characteristics. Water quality and aquatic plants were both cited as impacting recreational assessments, although the most significant impacts were associated with poor clarity and high algae levels.

The 2001 NYSDEC Priority Waterbody Listings (PWL) for the Lower Hudson River drainage basin do not include Lake Oscaleta. The CSLAP datasets are not yet adequate to recommend a PWL listing for the lake. The next PWL review for the Lower Hudson River drainage basin will likely occur in 2008.

General Comments and Questions:

- ***What is the condition of Lake Ooscaleta?***

Water quality conditions in Lake Ooscaleta may not be adequate to support some recreational uses of the lake during the summer, and the lake can be best described as eutrophic, or highly productive. The extent of algae growth is strongly influenced by nutrients (perhaps phosphorus and nitrogen), and changes in algae are likely to impact water transparency. Recreational assessments are impacted by poor water clarity and rooted aquatic plant growth. However, plant communities have not been evaluated through CSLAP.

- ***What about the dark and murky bottom waters of the lake?***

The deep waters of Lake Ooscaleta have higher phosphorus readings than those at the lake surface, and data from other programs indicate deepwater oxygen deficits. It is not yet known what impact these deepwater oxygen deficits (and nutrient-enriched) bottom waters has on surface water quality, particularly since seasonal water quality patterns were not apparent, at least in 2006.

- ***How does this condition change from spring showers thru changing of the leaves?***

The productivity of Lake Ooscaleta (as measured by clarity, nutrient and algae levels) appears to vary randomly during the summer, despite what may be a seasonal migration of nutrients from bottom to surface waters. Additional data will help to determine if these seasonal “trends” consistently occur in the lake.

- ***How has the condition changed since CSLAP sampling began on the lake and/or relative to historical values?***

2006 was the first year of CSLAP sampling at Lake Ooscaleta. Historical (1987 and 2003) data suggest that the lake may now be more productive, based on lower water transparency and higher nutrient, algae, pH, and conductivity readings. Additional data will help to document any long-term trends.

- ***How does Lake Ooscaleta compare to other similar lakes (nearby lakes,....)?***

Lake Ooscaleta appears to be more productive than other lakes classified for contact recreation (Class B), other nearby (Lower Hudson River drainage basin) lakes, and other NYS lakes. However, recreational assessments in Lake Ooscaleta are more favorable than in these other lakes, perhaps suggesting that these represent “normal” conditions or non-water quality factors (such as weeds) exert less impact.

- ***Based on these data, what should be done to improve or maintain Lake Ooscaleta?***

It is premature to identify lake management recommendations for the lake, given the relative lack of water quality data. However, as with most NYS lakes, it is likely that the management of water quality conditions in Lake Ooscaleta should focus on reducing nutrient and sediment loading to the lake, through pumping and maintaining septic systems, utilizing shoreline buffer zones, limiting use of lawn fertilizers, minimizing land disturbances in the near-lake watershed, and localized stormwater management. The lake association is also advised to minimize introductions of exotic plants and animals into the lake, particularly given the strong connection between weeds and recreational assessments of the lake, and the increasing presence of invasive exotic plants in southern New York State.

Context and Qualifiers

The NY Citizens Statewide Lake Assessment Program (CSLAP) is intended to be a long-term, standardized, trophic-based water quality monitoring program to facilitate comparison of water quality data from season to season, year to year, and from lake to lake. The data and information collected through CSLAP can be utilized to identify water quality problems, detect seasonal and long-term patterns, and educate sampling volunteers and lake residents about water quality conditions and stressors at their lakes. It is particularly useful in evaluating the over-enrichment of aquatic plant (algae and rooted plant) communities in a lake, and the response of the lake to these trophic stressors.

Shorefront residents, lake managers, and government agencies are increasingly tasked to better assess and evaluate water quality conditions and lake uses in NYS lakes, including those sampled through CSLAP, whether to address localized problems, meet water quality standards, satisfy state and federal environmental reporting requirements, or enhance and balance a suite of lake uses. CSLAP data should be a part of this process, but only a part. For some lakes, particularly small lakes and ponds with limited public access by those who don't reside on the lake shore, CSLAP may be the sole source of data used to assess lake conditions. In addition, studies conducted through CSLAP find strong similarities between sampling sites in many, but not all, large lakes, and generally find a strong convergence of perceptions about lake and recreational use conditions within most lakes, based on a local familiarity with "normal" conditions and factors that might affect lake use. For the purpose of broad water quality evaluations and understanding the connection between measured water quality indicators and the support of broadly-based recreational uses of the lake, CSLAP can be a singularly effective tool for standardizing the lake assessment process. CSLAP volunteers, lake associations, and others engaged in lake assessment and management should continue to utilize CSLAP in this context.

However, for large, multi-use lakes, or those lakes that are threatened by pollutants not captured in eutrophication-based monitoring programs, CSLAP becomes a less effective primary tool for assessing lake condition and use impairments. For example, CSLAP data have only limited utility in evaluating the following:

- (a) contamination from bacteria or other biological toxins, particularly related to the safety of water use for potable intake or swimming
- (b) contamination from inorganic (e.g., metals) and organic (e.g., PCBs, DDT) compounds
- (c) portions of a lake not well-mixed with the "open water" or otherwise distant from the primary sampling site(s), including the shoreline, bottom sediment and isolated coves
- (d) rooted aquatic plant impacts in areas of the lake not evaluated by the sampling volunteers
- (e) diverging perceptions of recreational use impacts, particularly in lakes with shorelines or isolated coves exhibiting conditions very different from those sampled or evaluated by the sampling volunteers
- (f) impacts to fish or other fauna due to factors unrelated to eutrophication
- (g) PWL or 303(d) listings for other pollutants or portions of the lake not sampled through CSLAP

For these waterbodies, CSLAP can and should continue to be part of an extensive database used to comprehensively evaluate the entirety of the lake and its uses, but absent a more complete dataset, CSLAP data should be used with caution as a sole means for evaluating the lake. Water quality evaluations, recommended PWL listings, and other extrapolations of the data and analyses should be utilized in this context, and by no means should be considered "the last word" on the lake.

I. INTRODUCTION: CSLAP DATA AND YOUR LAKE

Lakes are dynamic and complex ecosystems. They contain a variety of aquatic plants and animals that interact and live with each other in their aquatic setting. As water quality changes, so too will the plants and animals that live there, and these changes in the food web also may affect water quality. Water quality monitoring provides a window into the numerous and complex interactions of lakes. Even the most extensive and expensive monitoring program **cannot completely assess** the water quality of a lake. However, by looking at some basic chemical, physical, and biological properties, it is possible to gain a greater understanding of the general condition of lakes. CSLAP monitoring is a basic step in overall water quality monitoring.

Understanding Trophic States

All lakes and ponds undergo **eutrophication**, an aging process, which involves stages of succession in biological productivity and water quality (see Figure 1). **Limnologists** (scientists who study fresh water systems) divide these stages into **trophic** states. Each trophic state can represent a wide range of biological, physical, and chemical characteristics and any lake may “naturally” be categorized within any of these trophic states. In general, the increase in productivity and decrease in clarity corresponds to an enrichment of nutrients, plant and animal life. Lakes with low biological productivity and high clarity are considered **oligotrophic**. Highly productive lakes with low clarity are considered **eutrophic**. Lakes that are **mesotrophic** have intermediate or moderate productivity and clarity. It is important to remember that eutrophication is a natural process, and is not necessarily indicative of man-made pollution.

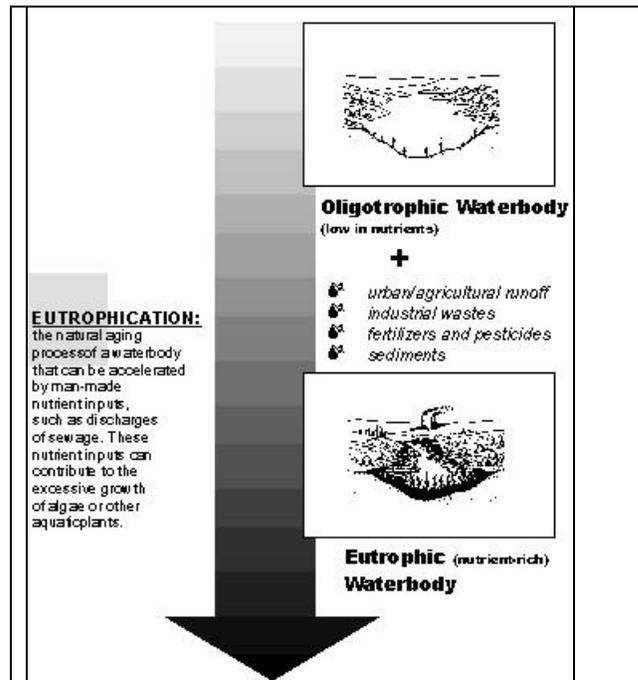


Figure 1. Trophic States

In fact, some lakes are thought to be “naturally” productive. Trophic classifications are not interchangeable with assessments of water quality. Water quality degradation from the perspective of one user may contrast with the perception of favorable conditions by a different lake user. For example, a eutrophic lake may support an excellent warm-water fishery because it is nutrient rich, but a swimmer may describe that same lake as polluted. A lake’s trophic state is still important because it provides lake managers with a reference point to view changes in a lake’s water quality and begin to understand how these changes may cause **use impairments** (threaten the use of a lake or swimming, drinking water or fishing).

When human activities accelerate lake eutrophication, it is referred to as **cultural eutrophication**. Cultural eutrophication may result from shoreline erosion, agricultural and urban runoff, wastewater discharges or septic seepage, and other nonpoint source pollution sources. These can greatly accelerate the natural aging process of lakes, cause successional changes in the plant and animal life within the lake, shoreline and surrounding watershed, and impair the water quality and value of a lake. They may ultimately extend aquatic plants and emergent vegetation throughout the lake, resulting in the transformation of the lake into a marsh, prairie, and forest. The extent of cultural eutrophication, and the corresponding pollution problems, can be signaled by significant changes in the trophic state over a short period of time.

Why is this important? New York State lakes can be affected by a variety of stressors, from acid rain to zebra mussels and almost everything in between. In any given part of the state, some of these stressors are more important than others. For example, there are probably more lakes affected by acid rain than any other pollutant, but these impacts are typically associated with a particular region (the Adirondacks and Catskills) and particular type of lake (small, high elevation lakes in basins with thin soils and little buffering capacity). But for most lakes in New York, cultural eutrophication represents the most significant source of pollutants and threat to water quality. As a result, water quality indicators related to eutrophication comprise the foundation of most water quality monitoring programs.

II. CSLAP SAMPLING PARAMETERS

CSLAP monitors several parameters related to the trophic state of a lake, including the clarity of the water, the amount of nutrients in the water, and the amount of algae resulting from those nutrients. Three parameters are the most important measures of eutrophication in most New York lakes: **total phosphorus**, **chlorophyll *a*** (estimating the amount of algae), and **Secchi disk transparency**. Because these parameters are closely linked to the growth of weeds and algae, they provide insight into “how the lake looks” and its suitability for recreation and aesthetics. Other CSLAP parameters help characterize water quality at the lake. Each of these sampling parameters are outlined in Figure 2. In addition, CSLAP also uses the responses on the **Field Observation Forms** to gauge volunteer perceptions of lake water quality. Most water quality “problems” arise from impairment of accepted or desired lake uses, or the perception that such uses are somehow degraded. As such, any water quality monitoring program should attempt to understand the link between perception and measurable quality.

The parameters analyzed in CSLAP provide valuable information for characterizing lakes. By adhering to a consistent sampling protocol provided in the [CSLAP Sampling Protocol](#), sampling volunteers collect and use data to assess both seasonal and yearly fluctuations in these parameters, and to evaluate the water quality conditions in their lake. By comparing a specific year's data to historical water quality information, lake managers can pinpoint trends and determine if water quality is improving, degrading or remaining stable. Such a determination answers a first critical question posed in the lake management process.

Figure 2. CSLAP Parameters

PARAMETER	SIGNIFICANCE
Water Temperature (°C)	Water temperature affects many lake activities, including the rate of biological growth and the amount of dissolved oxygen. It also affects the length of the recreational season
Secchi Disk Transparency (m)	Determined by measuring the depth at which a black and white disk disappears from sight, the Secchi disk transparency estimates the clarity of the water. In lakes with low color and rooted macrophyte ("weed") levels, it is related to algal productivity
Conductivity (µmho/cm)	Specific conductance measures the electrical current that passes through water, and is used to estimate the number of ions (charged particles). It is somewhat related to both the hardness and alkalinity (acid-buffering capacity) of the water, and may influence the degree to which nutrients remain in the water. Generally, lakes with conductivity less than 100 µmho/cm are considered softwater, while conductivity readings above 300 µmho/cm are found in hardwater lakes.
pH	pH is a measure of the (free) hydrogen ion concentration in solution. Most clearwater lakes must maintain a pH between 6 and 9 to support most types of plant and animal life. Low pH waters (<7) are acidic, while high pH waters (>7) are basic
Color (true) (platinum color units)	The color of dissolved materials in water usually consists of organic matter, such as decaying macrophytes or other vegetation. It is not necessarily indicative of water quality, but may significantly influence water transparency or algae growth. Color in excess of 30 ptu indicate sufficient quantities of dissolved organic matter to affect clarity by imparting a tannic color to the water.
Phosphorus (total, mg/l)	Phosphorus is one of the major nutrients needed for plant growth. It is often considered the "limiting" nutrient in NYS lakes, for biological productivity is often limited if phosphorus inputs are limited. Nitrogen to phosphorus ratios of >10 generally indicate phosphorus limitation. Many lake management plans are centered on phosphorus controls. It is measured as total phosphorus (TP)
Nitrogen (nitrate, ammonia, and total (dissolved), mg/l)	Nitrogen is another nutrient necessary for plant growth, and can act as a limiting nutrient in some lakes, particularly in the spring and early summer. Nitrogen to phosphorus ratios <10 generally indicate nitrogen limitation (for algae growth). For much of the sampling season, many CSLAP lakes have very low or undetectable levels of one or more forms of nitrogen. It is measured in CSLAP in three forms - nitrate/nitrite (NO _x), ammonia (NH _{3/4}), and total nitrogen (TN or TDN).
Chlorophyll <i>a</i> (µg/l)	The measurement of chlorophyll <i>a</i> , the primary photosynthetic pigment found in green plants, provides an estimate of phytoplankton (algal) productivity, which may be strongly influenced by phosphorus
Calcium (mg/l)	Calcium is a required nutrient for most aquatic fauna, and is required for the shell growth for zebra mussels (at least 8-10 mg/l) and other aquatic organisms. It is naturally contributed to lakes from limestone deposits and is often strongly correlated with lake buffering capacity and conductivity.

Ranges for Parameters Assessing Trophic Status and Lake Osaleta

The relationship between phosphorus, chlorophyll *a*, and Secchi disk transparency has been explored by many researchers, to assess the trophic status (the degree of eutrophication) of lakes. Figure 3 shows ranges for phosphorus, chlorophyll *a*, and Secchi disk transparency (summer median) are representative for the major trophic classifications:

These classifications are valid for clear-water lakes only (with less than 30 platinum color units).

Some humic or “tea color” lakes, for example, naturally have high levels of dissolved organic material, resulting in color readings that exceed 30 color units.

Figure 3. Trophic Status Indicators

Parameter	Eutrophic	Mesotrophic	Oligotrophic	Lake Osaleta
Phosphorus (mg/l)	> 0.020	0.010 - 0.020	< 0.010	0.031
Chlorophyll <i>a</i> (µg/l)	> 8	2- 8	< 2	14.6
Secchi Disk Clarity (m)	< 2	2- 5	> 5	2.1

This will cause the water transparency to be lower than expected given low phosphorus and chlorophyll *a* levels in the lake. Water transparency can also be unexpectedly lower in shallow lakes, due to influences from the bottom (or the inability to measure the maximum water clarity due to the visibility

of the Secchi disk on the lake bottom). Even shallow lakes with high water clarity, low nutrient concentrations, and little algal growth may also have significant weed growth due to shallow water conditions. While such a lake may be considered unproductive by most water quality standards, that same lake may experience severe aesthetic problems and recreational impairment related to weeds, not trophic state. Generally, however, the trophic relationships described above can be used as an accurate "first" gauge of productivity and overall water quality.

By the total phosphorus and chlorophyll *a* trophic standards described above, Lake Oscaleta would probably be considered **eutrophic, or highly productive**, while by the Secchi disk transparency standards, the lake would be considered **mesotrophic, or moderately productive**. The most appropriate trophic designation for the lake is probably mesoeutrophic, or moderately to highly productive. The trophic condition of Lake Oscaleta is discussed in much greater detail below.

III. CSLAP LAKES

CSLAP sampling began in 1986 on 25 lakes generally distributed throughout the state, and in the following 20 years has expanded to more than 200 lakes. The program was developed primarily to identify water quality problems, develop long-term databases, and educate lakefront property owners on small lakes with little historical information and few other contemporary studies. However, the program has been utilized by lake residents, lake associations and managers, municipalities, state and federal government and environmental organizations to gain insights about small ponds, large high profile lakes and multi-use reservoirs from eastern Long Island to the northern Adirondacks to the western border of New York State. A map showing each of the lakes sampled through CSLAP since 1986 is shown in

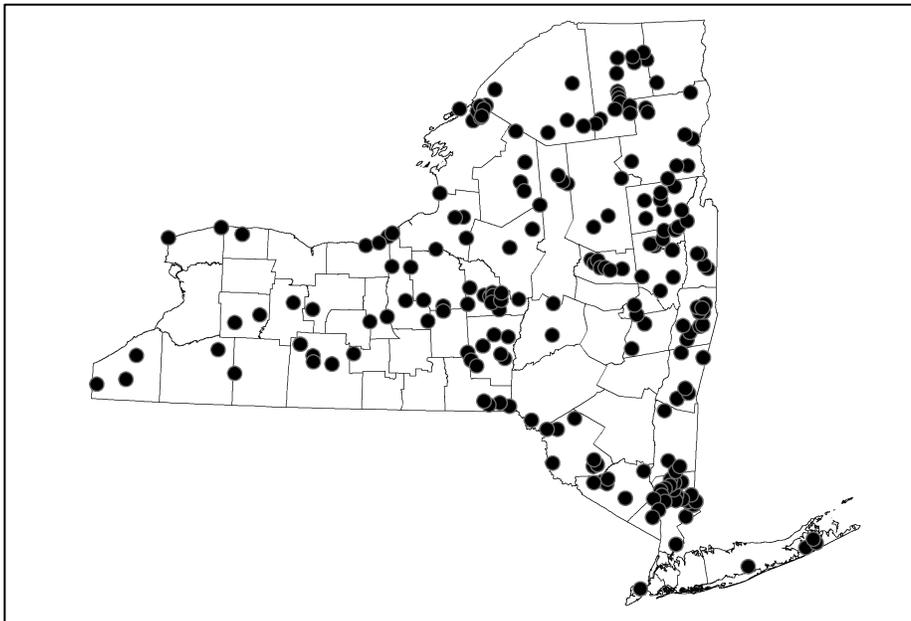


Figure 4. The distribution of lakes roughly matches the distribution of lake associations in the state (or at least those affiliated with the NY Federation of Lake Associations, the largest lake association organization in the state). The relative paucity of CSLAP lakes in the Finger Lakes region reflects the small number of lakes in a region dominated by very large lakes, while the small number of lakes sampled in the Catskills, Long Island, and Western NY reflects the shortage of organized lake associations in those areas.

Figure 4: CSLAP Sampling Sites

CSLAP lakes have ranged from the very small (5 acre Cranberry Lake in the downstate region) to the Great (two state park beaches on Lake Ontario). It has included perhaps the clearest lake in New York State (Skaneateles Lake, one of the Finger Lakes, with as high as 50 feet of water transparency) and several lakes with clarity as low as 1 foot. There are a large number of lakes used for potable water,

as well as those classified only for fishing and non-contact recreation. Some lakes (those on Long Island) sit just above sea level, while others are perched high in the clouds, including Summit Lake in central NY and Twitchell Lake in the Adirondacks, more than 2000 feet above sea level.

Figures 5a through 5d summarize the variety of lakes sampled through CSLAP. In short, these lakes constitute a comprehensive cross-section of the lake conditions, uses, and settings encountered in New York State.

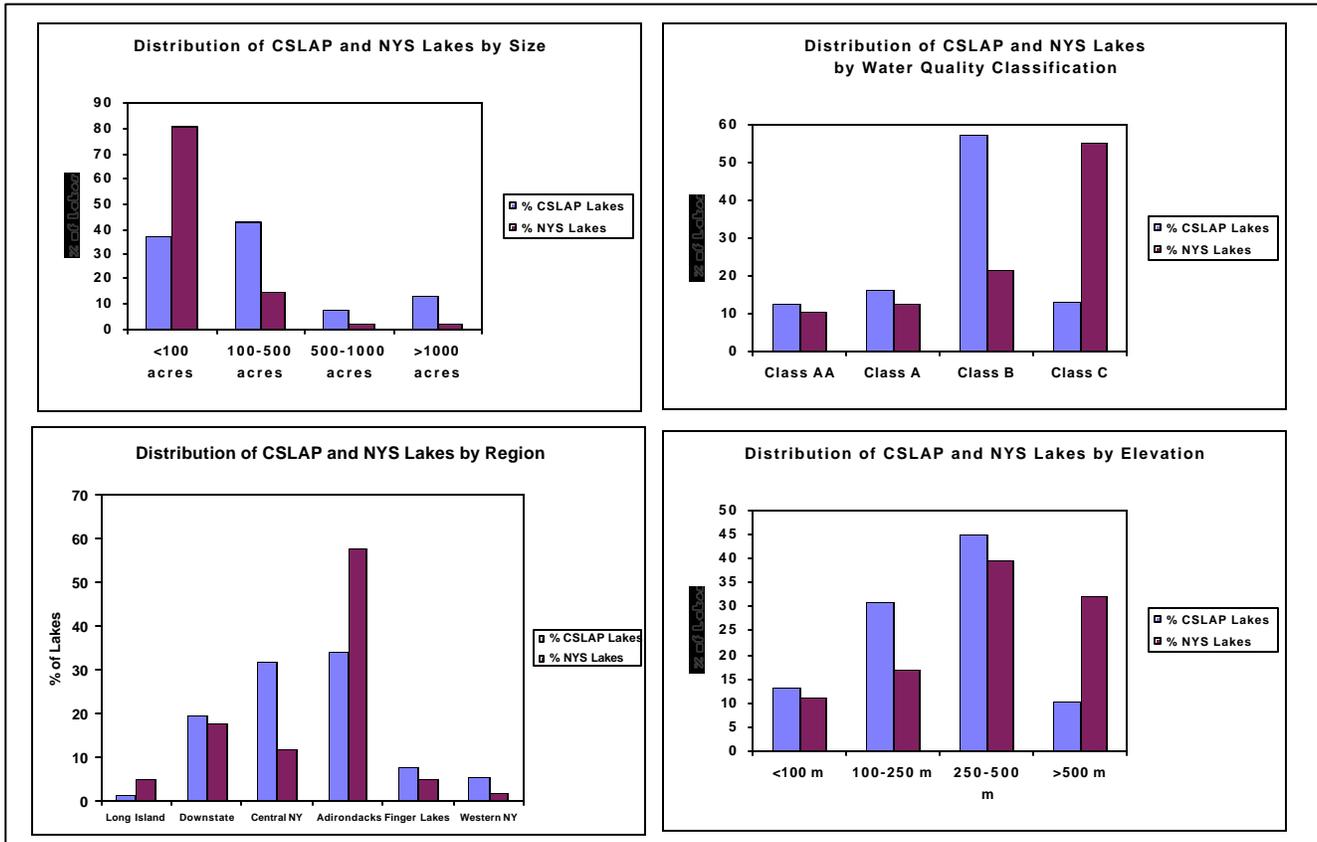


Figure 5- Comparison of CSLAP and New York State Lakes

The typical CSLAP lake is slightly larger than the typical New York State lake, and is more likely to be found in Central New York (between the Adirondacks, Downstate and Finger Lakes regions). However, this profile, as well as the preponderance toward “mid-elevation” regions, is probably more typical of the “lake community” regions of the state- those regions in which large numbers of lakes are heavily populated, which in turn represents lower elevation waterbodies that support siting septic systems, and have close proximity to roads and other non-lake communities (comprised of visitors and seasonal lake residents). The relatively higher percentage of Class B lakes in CSLAP and Class C lakes in the rest of the state reflects the large number of uninhabited Class C lakes in the Adirondacks. These lakes have been classified as Class C lakes, often by default, due in part to the lack of information about historical or contemporary lake uses and water quality conditions.

The distribution of lakes in these categories does suggest that CSLAP lakes are mostly comparable to other New York State lakes, and that an evaluation of CSLAP data may serve as a reasonable surrogate for statewide water quality evaluations, particularly since CSLAP serves as the primary long-term database maintained and supported by New York State.

IV: LAKE OSCALETA- BACKGROUND INFORMATION

Lake Oscaleta is a 65 acre lake found in the town of Lewisboro in Westchester County, just north of the New York City region of New York State. Figure 6 shows the location of Lake Oscaleta. It is one of 15 CSLAP lakes among the >120 lakes found in Westchester County, and one of 41 CSLAP lakes among the >350 lakes and ponds in the Lower Hudson River drainage basin. Lake Oscaleta is a Class B lake- this means that the best intended use for the lake is for contact recreation, including swimming and bathing. These “categories” will be used to evaluate water quality conditions later in the report.

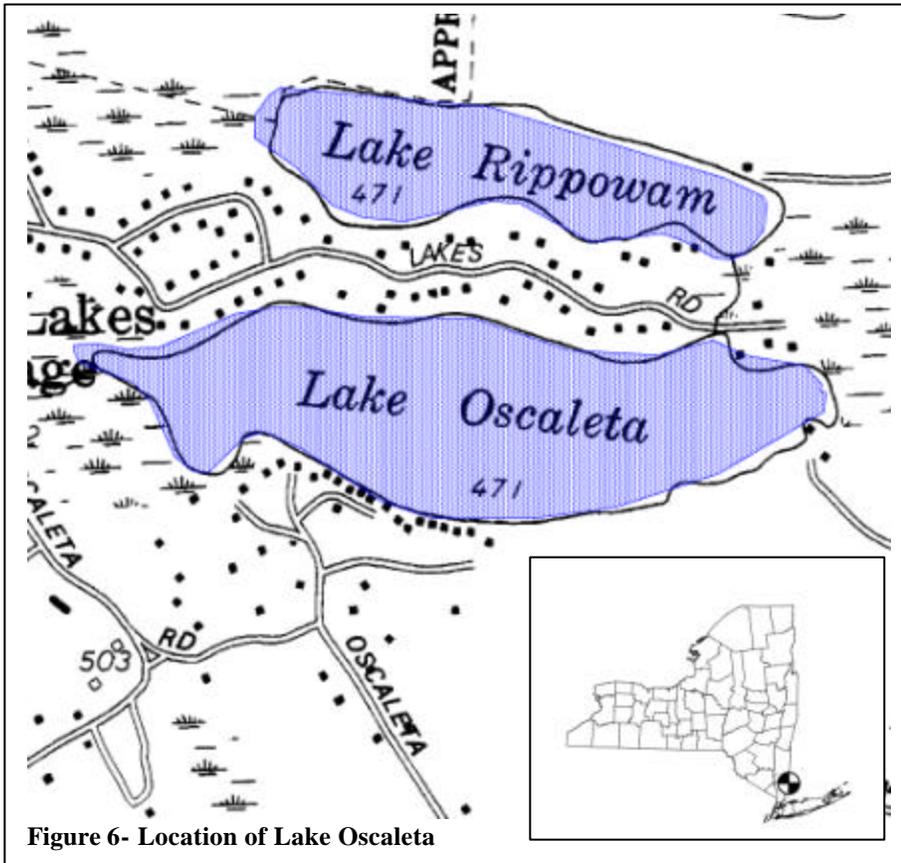


Figure 6- Location of Lake Oscaleta

CSLAP samples were collected from the deepest part of the lake, at least as determined by the CSLAP sampling volunteers. While it appears that the lake is thermally stratified, deepwater samples were not collected in 2006.

Historical Water Quality Information for Lake Oscaleta

Lake Oscaleta was sampled in 1987 as part of the Adirondack Lake Survey Corporation (ALSC) study of >1500 lakes in the Adirondacks and southern New York. These data are presented in Table 1, and showed that Lake Oscaleta was less productive in 1987, based on slightly higher water transparency and lower nutrient levels. Dissolved oxygen

readings were depressed near the lake bottom, although these readings did not bottom out. The lake was dominated by submergent plants- although milfoil and pondweeds were identified, it was not reported if these corresponded to exotic species. Phragmites were found at the lake. The lake was also sampled by Cedar Eden Environmental LLC in 2003 in anticipation of developing a Lake and Watershed Management Plan for the lake. These data, generally transcribed from graphs within the issued report, are also provided in Table 1.

It is not known if local monitoring has been conducted as a fisheries management tool, or to evaluate swimming conditions in the lake.

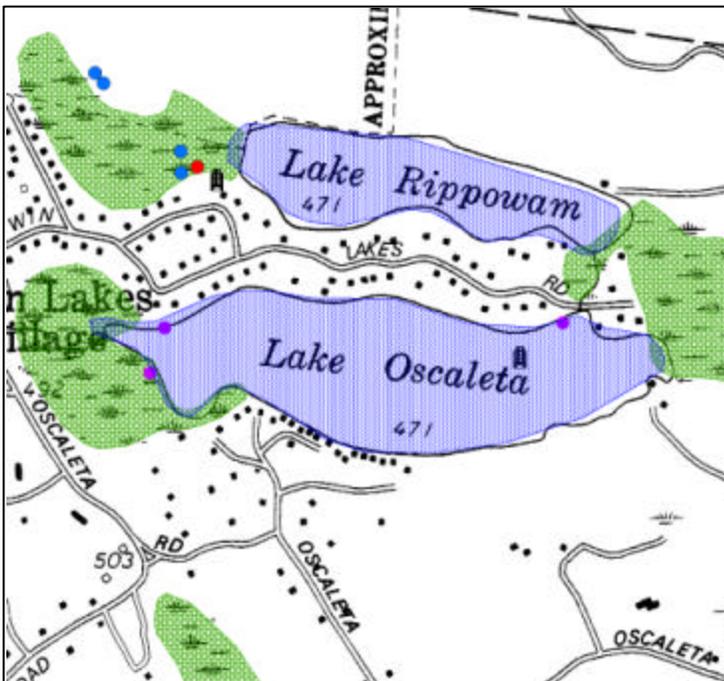
Historical Fisheries Information for Lake Oscaleta

It is not known if Lake Oscaleta has been stocked through any state fisheries stocking programs. The following local stocking record is reported by the Three Lakes Council for Lake Waccabuc, Lake Rippowam, and Lake Oscaleta:

1996 -no stocking
1997- ?
1998- 500 small mouth bass, 250 each in Rippowam and Oscaleta
1999- 1000 11" brown trout, 800 in Waccabuc, 200 in Oscaleta
2000- 500 11" brown trout
2001- 420 11" brown trout, 300 in Waccabuc, 120 in Oscaleta
2002- no stocking
2003- 500 10" brown trout, 350 in Waccabuc, 150 in Oscaleta
2004- 650 11" brown trout, 400 in Waccabuc, 250 in Oscaleta
2005- ?
2006- 650 11" brown trout, 425 in Waccabuc and 225 in Oscaleta

Prize white perch are found in the lake, with the state record white perch caught from the lake in 1991. Other fish species, as of 1987 (as noted in the ALSA survey of the lake), included alewife, bluegill, brown bullhead, chain pickerel, common carp, golden shiner, largemouth bass, pumpkinseed, white sucker, and yellow perch.

General statewide fishing regulations are applicable in Lake Oscaleta.



Permitted Facilities Associated with Lake Oscaleta

There appear to be only a small number of facilities on or near Lake Oscaleta that require permits or are otherwise regulated by the NYSDEC. These correspond to private residences (or wells just to the west of Lake Rippowam), and are shown as "milkcan" symbols on the map on the left. The green crosshatched areas correspond to regulated wetlands.

V. NEW YORK STATE, CSLAP AND LAKE OSCALETA WATER QUALITY DATA: 1986-2005

Overall Summary:

Although water quality conditions at each CSLAP lake have varied each year since 1986, and although detailed statistical analyses of the entire CSLAP dataset has not yet been conducted, general water quality trends can be evaluated after 5-20 years worth of CSLAP data from these lakes. Overall (regional and statewide) water quality conditions and trends can be evaluated by a variety of different means. Each of the tested parameters (“analytes”) can be evaluated by looking at the how the analyte varies from year to year from the long-term average (“normal”) condition for each lake, and by comparing these parameters across a variety of categories, such across regions of the state, across seasons (or months within a few seasons), and across designated best uses for these lakes. Such evaluations are provided in the second part of this summary, via Figures 7 through 17. The annual variability is expressed as the difference in the annual average (mean) from both the long-term average and the normal variability expected from this long-term average. The latter can be presented as the “standard error” (SE- calculated here within the 95% confidence interval) - one standard error away from the long-term average can be considered a “moderate” change from “normal”, with a deviation of two or more standard errors considered to be a “significant” change. For each of these parameters, the percentage of lakes with annual data falling within one standard error from the long-term average are considered to exhibit “no change”, with the percentage of lakes demonstrating moderate to significant changes also displayed on these graphs (Figures 7a through 17a). Annual changes in these lakes can also be evaluated by standard linear regressions- annual means over time, with moderate correlation defined as $R^2 > 0.33$, and significant correlation defined as $R^2 > 0.5$. These methods are described in greater detail in Appendix D. Assessments of weather patterns- whether a given year was wetter or drier than usual- accounts for broad statewide patterns, not weather conditions at any particular CSLAP lake. As such, weather may have very different at some (but not most) CSLAP lakes in some of these years.

Long-term trends can also be evaluated by looking at the summary findings of individual lakes, and attempting to extrapolate consistent findings to the rest of the lakes. Given the (non-Gaussian) distribution of many of the water quality parameters evaluated in this report, non-parametric tools may be the most effective means for assessing the presence of a water quality trend. However, these tools do not indicate the magnitude of the trend. As such, a combination of parametric and non-parametric tools are employed here to evaluate trends. The Kendall tau ranking coefficient has been utilized by several researchers and state water quality agencies to evaluate water quality trends via non-parametric analyses, and is utilized here. For parametric analyses, best-fit analysis of summer (June 15 through September 15) averages for each of the eutrophication indicators can be evaluated, with trends attributable to instances in which deviations in annual means exceed the deviations found in the calculation of any single annual mean. “Moderate” change is defined as $t > 0.33$, and “significant” change is defined as $t > 0.5$. It has been demonstrated in many of these programs that long-term trend analyses cannot be utilized to evaluate lake datasets until at least five years worth of data have been collected.

As of 2006, there were 112 CSLAP lakes sampled in the last five years that have been sampled for at least five years- the change in these lakes is demonstrated in Figures 7 and 8; Figures 7a through 7j indicate “moderate” long-term change, while Figures 8a through 8j indicate “significant” long-term change. When these lakes are analyzed by this combination of parametric and non-parametric analyses, these data suggest that while most NYS lakes have not demonstrated a significant change (either t or $R^2 > 0.5$) or even a moderate changes (t or $R^2 > 0.33$).

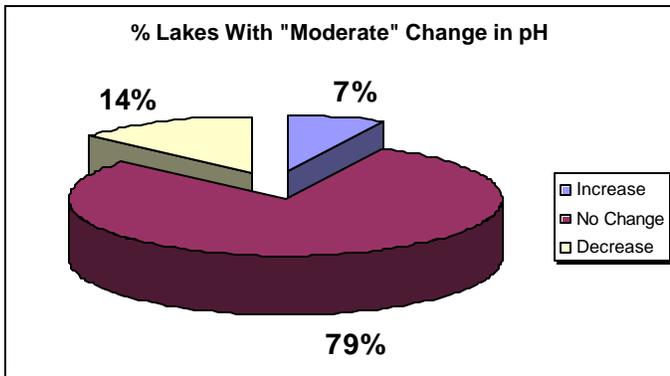


Figure 7a. %CSLAP Lakes Exhibiting Moderate Long-Term Change in pH

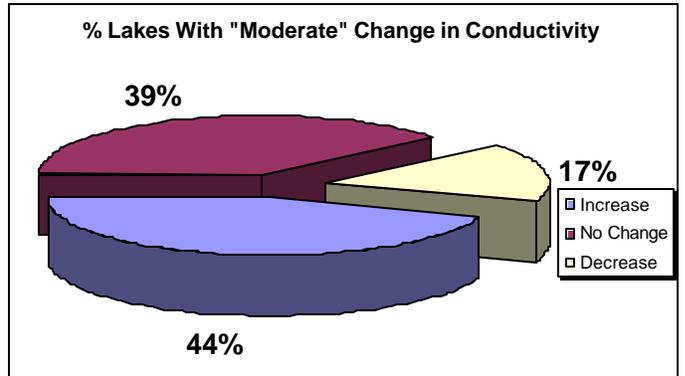


Figure 7b. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Conductivity

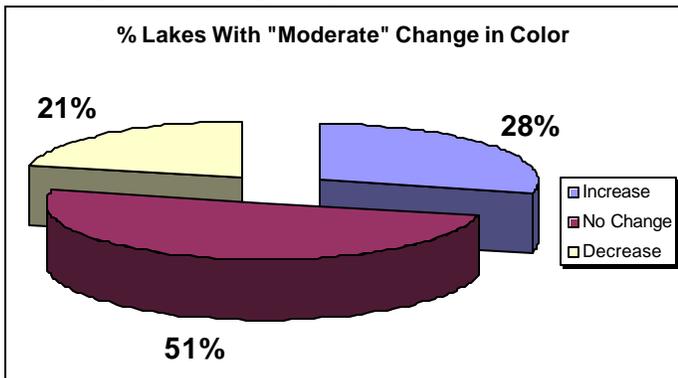


Figure 7c. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Color

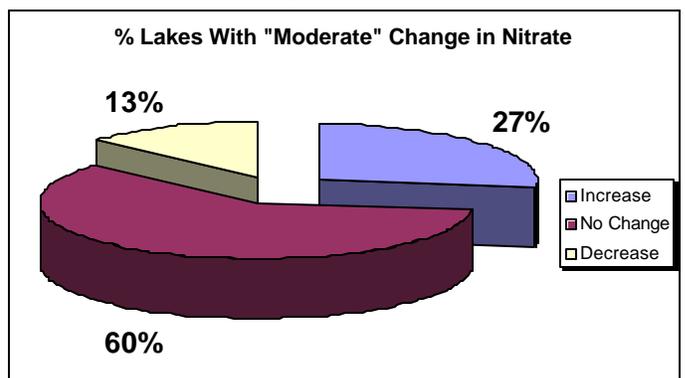


Figure 7d. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Nitrate

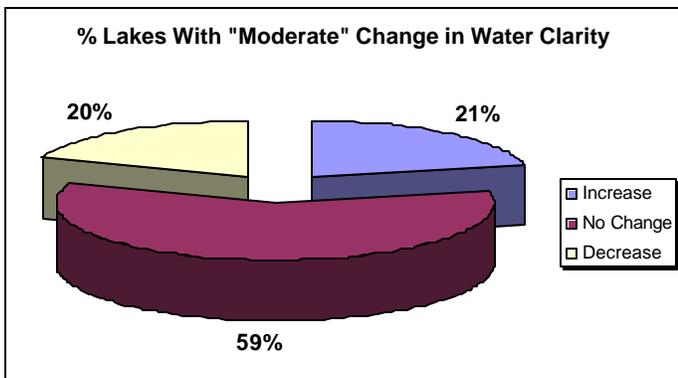


Figure 7e. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Water Clarity

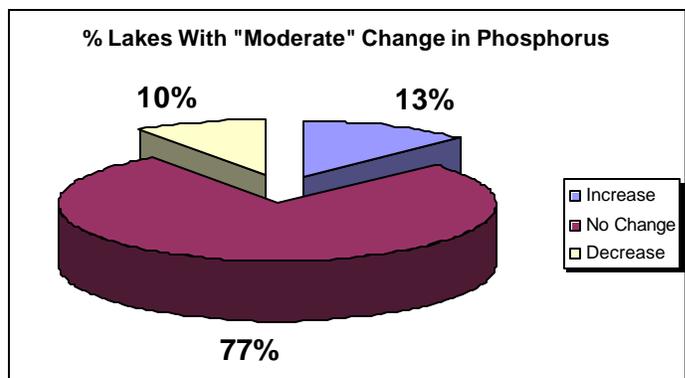


Figure 7f. %CSLAP Lakes Exhibiting Moderate Long-Term Changes in Phosphorus

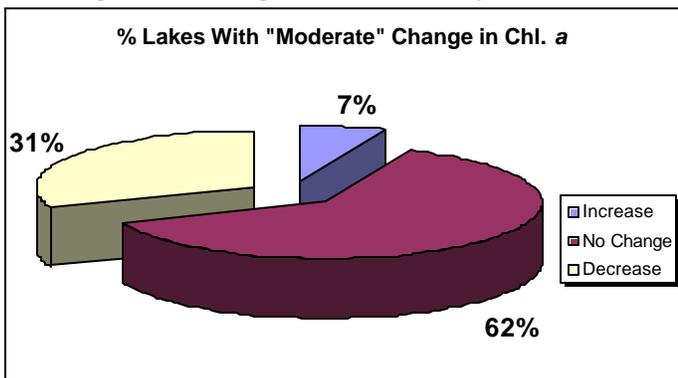


Figure 7g. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Chlorophyll a

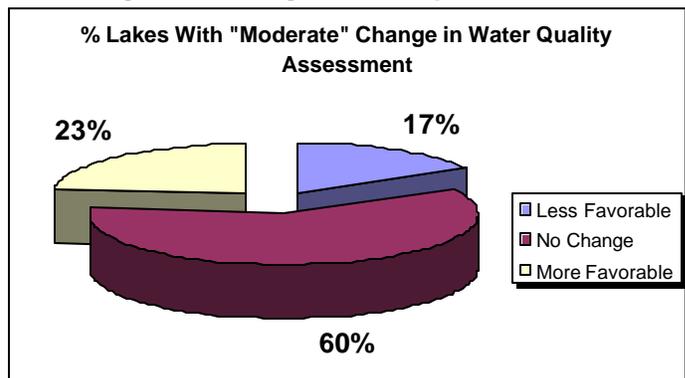


Figure 7h. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Water Quality Assessment

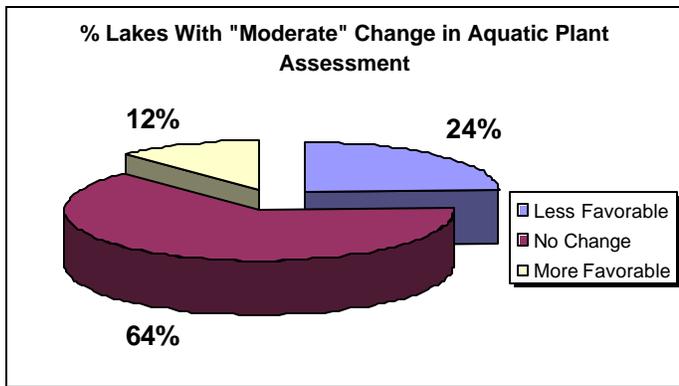


Figure 7i. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Aquatic Plant Assessment

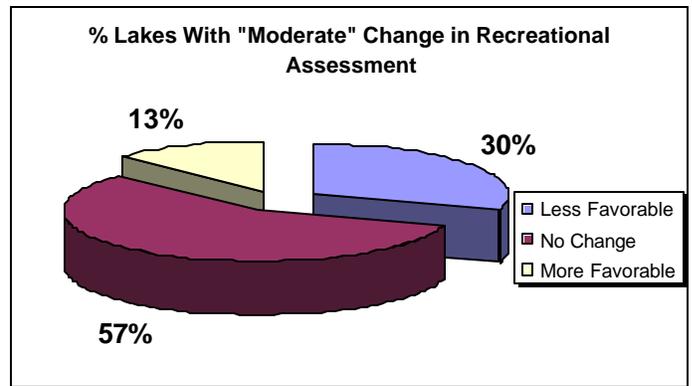


Figure 7j. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Recreational Assessment

Some of the lakes sampling through CSLAP have demonstrated a moderate change since CSLAP sampling began in 1986, at least for some of the sampling parameters measured through CSLAP. In general, between 50% and 65% of the CSLAP lakes have not exhibited even moderate changes. Some of the parameters that have exhibited moderate changes may not reflect actual water quality change. For example, it appears that the increase in color (Figure 7c) and decrease in nitrate (Figure 7d) and chlorophyll *a* (Figure 7g) is probably due to the shift in laboratories, even though the analytical methods are comparable. The increase in conductivity (Figure 7b) and decrease in pH (Figure 7a) are probably real phenomena- both changes were evident to some degree prior to the shift in laboratories, and both are largely predictable. The difference between the increase and decrease in the other sampling parameter (or between more favorable and less favorable conditions) does not appear to be important, and probably indicates random variability.

Figures 8a through 8j indicate that, not surprisingly, "substantial" change is less common. Substantial change follows the same patterns as discussed above with the evaluation of "moderate" change in CSLAP lakes, except that the percentage of CSLAP lakes not exhibiting significant change is much higher, rising to about 65-80% of these lakes. For those CSLAP lakes exhibiting substantial change, it is most apparent in the same parameters described above. About 25% of the CSLAP lakes have exhibited a substantial increase in conductivity, consistent with a broad (and expected) successional pattern, in which lakes generally concentrate materials washed in from the surrounding watershed (and as the runoff itself concentrates materials as these watersheds move from forested to more heavily used, whether via residential development or other uses. The comparison between Figures 8b and 8e through 8g indicate that this has not (yet) translated into higher nutrient loading into lakes.

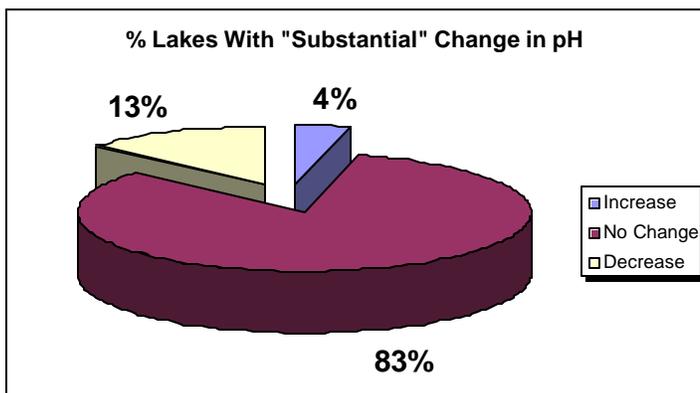


Figure 8a. %CSLAP Lakes Exhibiting Substantial Long-Term Change in pH

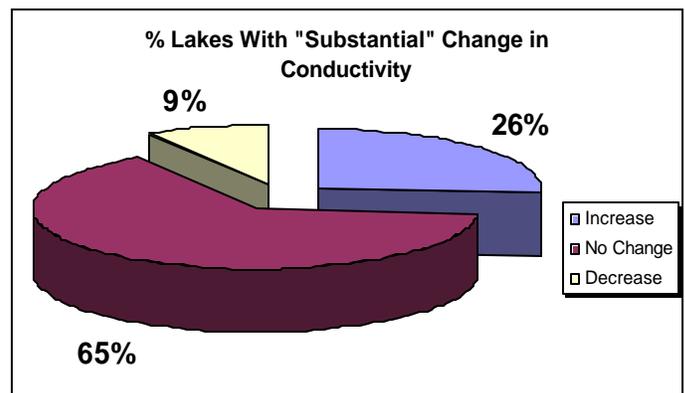


Figure 8b. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Conductivity

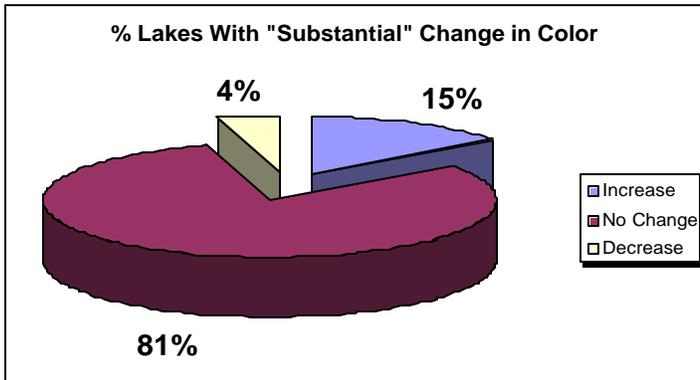


Figure 8c. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Color

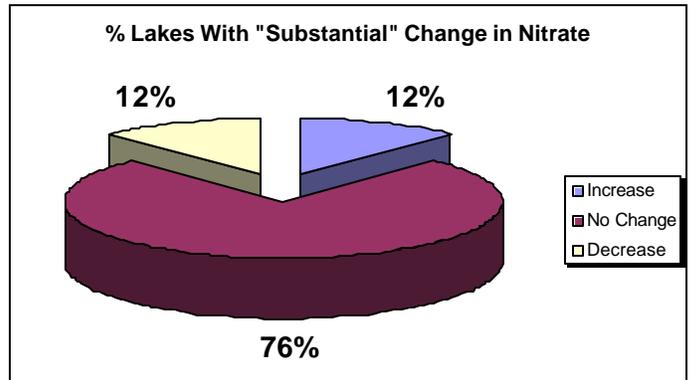


Figure 8d. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Nitrate

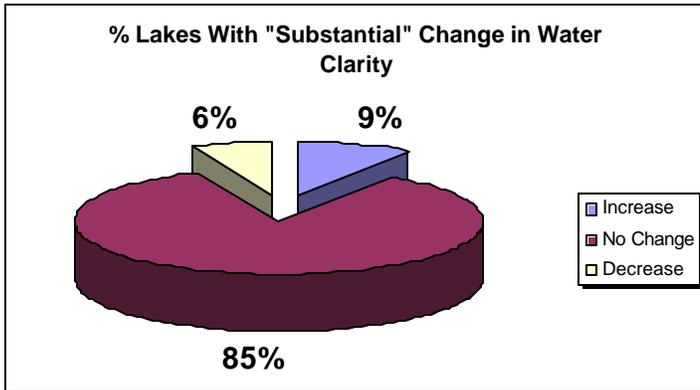


Figure 8e. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Water Clarity

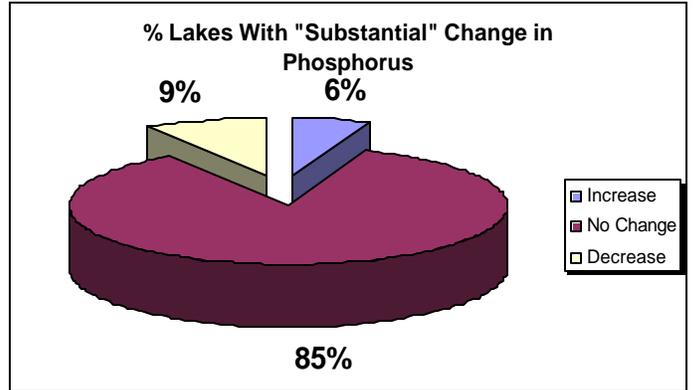


Figure 8f. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Phosphorus

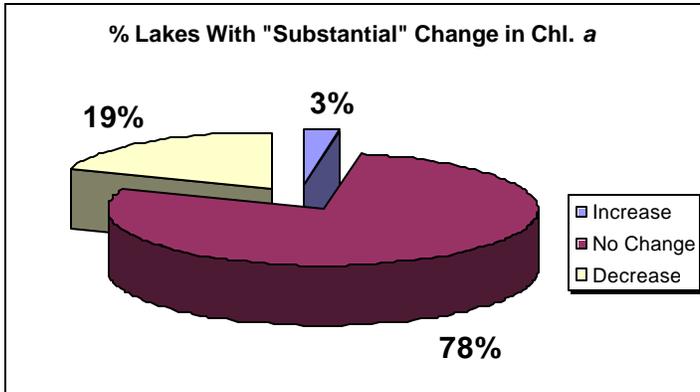


Figure 8g. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Chlorophyll a

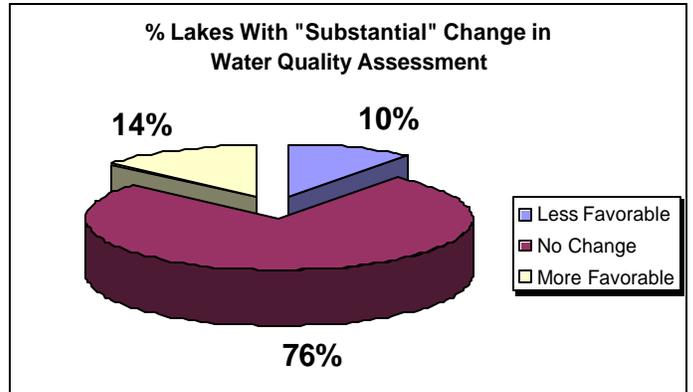


Figure 8h. %CSLAP Lakes Exhibiting Substantial Long-Term Changes in Water Quality Assessment

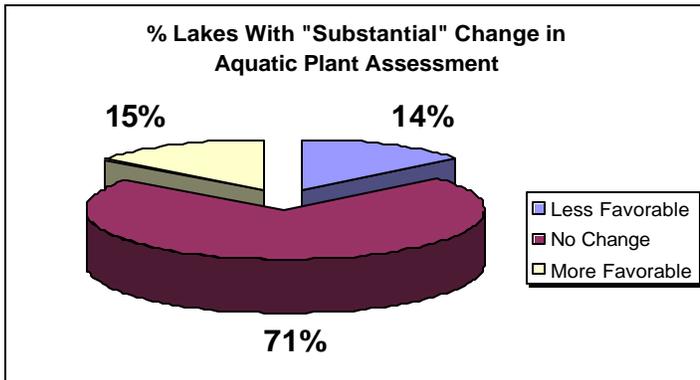


Figure 8i. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Aquatic Plant Assessment

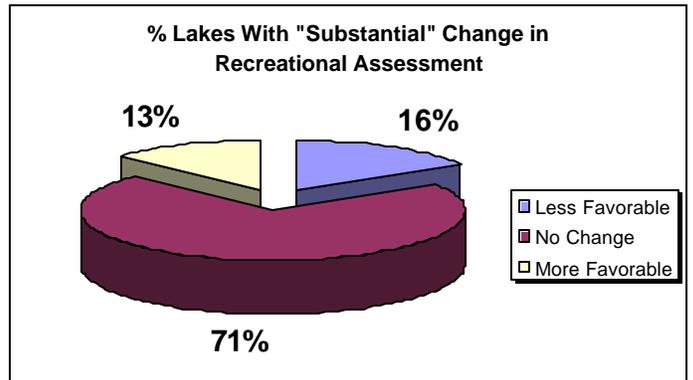


Figure 8j. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Recreational Assessment

As noted above, there does not appear to be any clear pattern between weather and water quality changes, although some connection between changes in precipitation and changes in some water quality indicators is at least alluded to in some cases. However, all of these lakes may be the long-term beneficiaries of the ban on phosphorus in detergents in the early 1970's, which with other local circumstances (perhaps locally more "favorable" weather, local stormwater or septic management, etc.) has resulted in less productive conditions. Without these circumstances, water quality conditions in many of these lakes might otherwise be more productive, in the creeping march toward aging, eutrophication, and succession (as suggested from the steady rise in conductivity). In other words, the higher materials loading into these lakes may be largely balanced by a reduction in nutrients within the corresponding runoff

The drop in pH in NYS lakes has been studied at length within the Adirondacks and may continue to be attributable on a statewide basis to acid rain, since acidic rain continues to fall throughout the state. The CSLAP dataset is not adequate to evaluate any ecological changes associated higher lake acidity, and it is certainly worth noting that the slight drop in pH in most CSLAP lakes does not bring these lakes into an acidic status (these lakes have, at worse, become slightly less basic). In addition, for lakes most susceptible to acidification, laboratory pH is only an approximation of actual pH. Fully accurate pH readings require field measurements using very specialized equipment, although for most lakes with even modest buffering capacity, laboratory pH is a good estimate of *in situ* pH readings. So while the decrease in pH in some CSLAP lakes should continue to be watched, it does not appear to be a cause for concern, at least relative to the low pH in small, undeveloped, high elevation lakes within the Adirondack Park.

Lake perception has changed more significantly than water quality (except conductivity), due in part to the shorter timeframe for evaluation and thus a lower statistical hurdle for quantifying change (15 years versus up to 20 years for some lakes), but perhaps due to the multiple influences of these phenomena. None of these indicators- water quality perception, weeds perception, or recreational perception- have varied in a consistent manner, although variability is more common in each of these indicators. The largest change is in recreational assessments, with about 1/3 of all lakes exhibiting substantial change and nearly half have demonstrated moderate change. A more detailed analysis of these assessments (not presented here) indicate that the Adirondacks have demonstrated more "positive" change than other regions of the state, due to the perception that aquatic weed densities have not increased as significantly (and water quality conditions have improved in some cases). However, the rapid spread of *Myriophyllum spicatum* into the interior Adirondacks will likely reverse this "trend" in coming years, and it is not clear if these "findings" can be extrapolated to other lakes within the Adirondack Park.

Larger trends and observations about each of the CSLAP sampling parameters are presented below in Figures 8 through 18. As noted in the nitrate discussion, there is still an insufficient database for ammonia or total nitrogen to evaluate annual, geographic, seasonal, or lake use variability in these sampling parameters. However, these parameters are discussed in the specific discussions for Lake Oseleta later in this document.

pH

Annual Variability

The pH of most CSLAP lakes has consistently been well within acceptable ranges for most aquatic organisms during each sampling season. The average pH has not varied significantly from one sampling season to the next. There does not appear to be a strong connection between pH and weather; some of the years with the relatively highest pH, 1988 and 1992, and the lowest pH, 1987, correspond to dry (1988), wet (1992), and normal (1987) years, although some of the other years with relatively low pH corresponded to wetter years (1996, 2000, and 2004). There does not appear to be any significant annual pH trends in the CSLAP dataset. 90% of all samples had pH between 6.5 and 8.5 (the state water quality standards); 6% of samples have pH > 8.5 and 4% have pH < 6.5.

What Was Expected in 2006?

2006 was a relatively wet year, at least in most of the state during much of the summer sampling season. While there is not a strong correlation between weather and pH during at least most of the CSLAP sampling seasons, pH readings have generally been lower during wet years, most likely to due the input of acidic rain. Therefore, it is anticipated that pH readings may be slightly lower than usual, at least in some CSLAP lakes.

And What Happened at Lake Oscaleta in 2006?

2006 was the first year of CSLAP sampling at Lake Oscaleta. A comparison of the 2006 data to the ALSC (1987) and Cedar Eden (2003) datasets indicate higher pH readings in 2006. This is probably consistent with the higher algae levels, since algal production will remove carbon dioxide (a weak acid) from the water during daily photosynthesis.

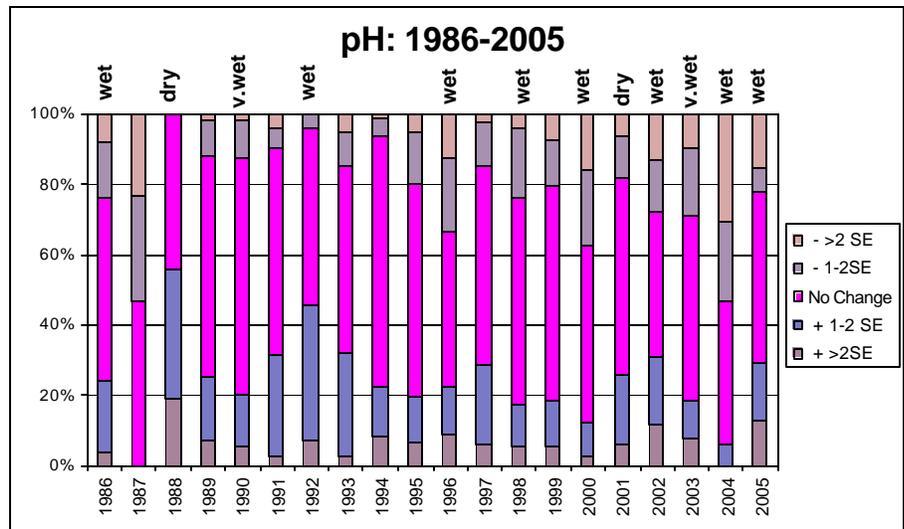


Figure 9a. Annual Change from "Normal" pH in CSLAP Lakes (SE = Standard Error)

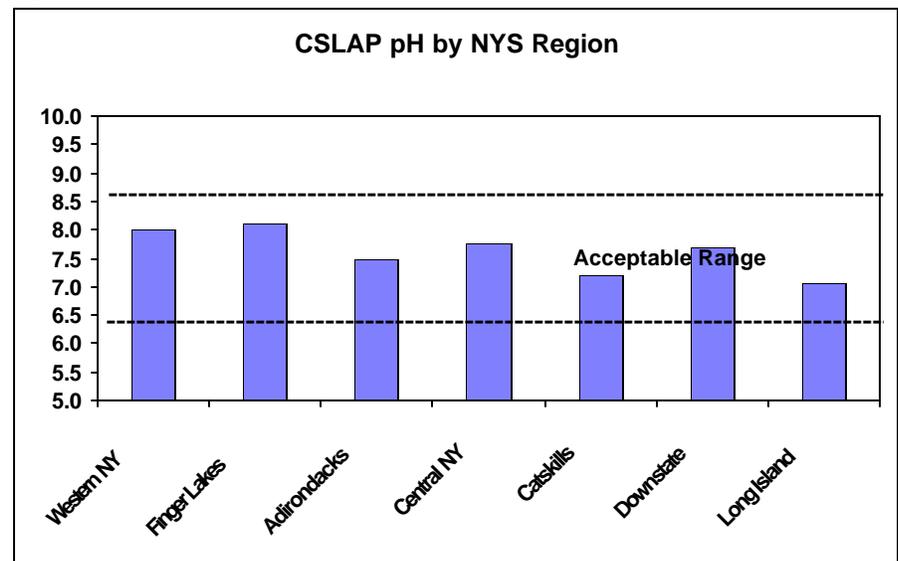


Figure 9b. pH in CSLAP Lakes by NYS Region

Statewide Variability:

As expected, pH readings are lowest in the high elevation regions (Adirondacks and Catskills) or Long Island, which has primarily shallow and slightly colored lakes, and highest in regions with relatively high conductivity (Western NY and the Finger Lakes region). All of these readings are consistently within the acceptable range for most aquatic organisms. However, the CSLAP dataset does not reflect the low pH found in many high elevation NYS lakes overlying granite and poorly buffered soils, since the typical CSLAP lake resides in geological settings (primarily limestone) that allow for residential development. In other words, pH is one of the few CSLAP sampling parameters that does not yield comparable results when comparing CSLAP results to overall NYS results, since CSLAP lakes are not really representative of the typical NYS lake as related to pH.

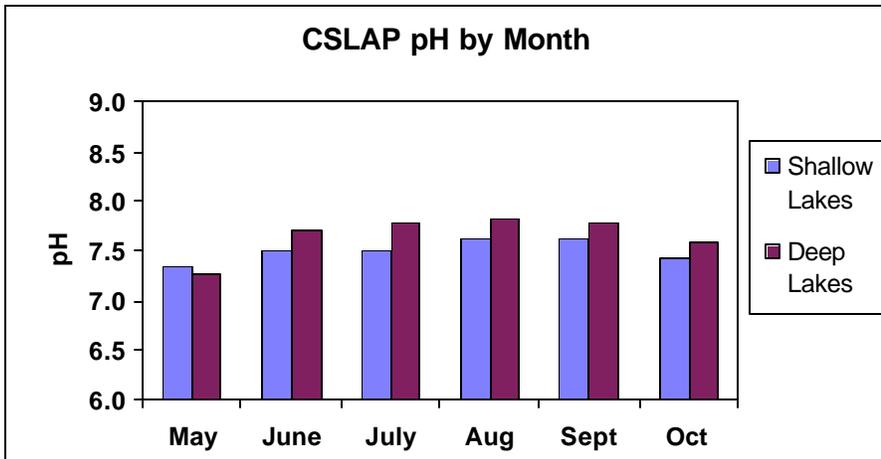


Figure 9c. pH in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

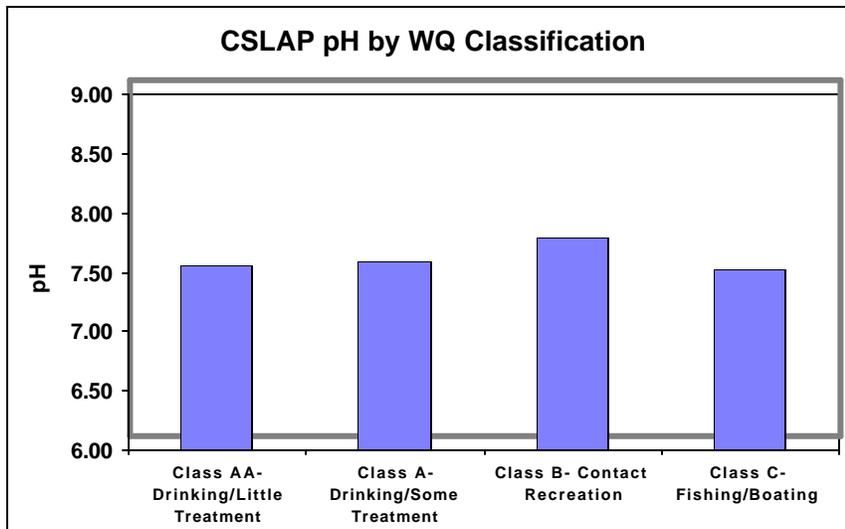


Figure 9d. pH in CSLAP Lakes by Lake Use

lake usage.

Seasonal Variability:

pH readings tend to increase slightly over the course of the summer, due largely to increasing algal photosynthesis (which consumes CO₂ and drives pH upward), although these seasonal changes are probably not significant. Low pH depressions are most common early in the sampling season (due to lingering effects from snowpack runoff) and high pH spikes occur mostly in mid to late summer.

Lake Use Variability

pH does not vary significant from one lake use to another, although in general pH readings are slightly higher for lakes used primarily for contact recreation (Class B). However, this is probably more reflective of geographical differences (there are relatively more Class B CSLAP lakes in higher pH regions, and more Class A lakes in lower pH regions) than any inherent link between pH and

Conductivity

Annual Variability

The conductivity of most CSLAP lakes has varied somewhat from year to year, and has been (slightly) increasing overall and in specific lakes since 1986. This is apparent from Figure 10a, which shows that more lakes have exhibited higher readings in recent years that in the first several years of CSLAP sampling at the lake (although lower conductivity was apparent in 2004). There does not appear to be a correlation between weather and conductivity, although the overall annual trend appears to be stronger than weather-impacted changes.

What Was Expected in 2006?

2006 was a relatively wet year, at least in most of the state during much of the summer sampling season. Conductivity readings have generally not been correlated with weather, and any weak weather patterns appear to be dwarfed with a significant trend toward increasing conductivity readings over time. Therefore, it is anticipated that conductivity readings may be within the normal range for most CSLAP lakes, although higher conductivity readings may be more likely than lower conductivity.

And What Happened at Lake Oscaleta in 2006?

2006 was the first year of CSLAP sampling at Lake Oscaleta. Conductivity readings in 2006 were higher than those measured in 1987. While this is a common trend in many NYS lakes (conductivity increasing over time), this may also be consistent with the higher pH.

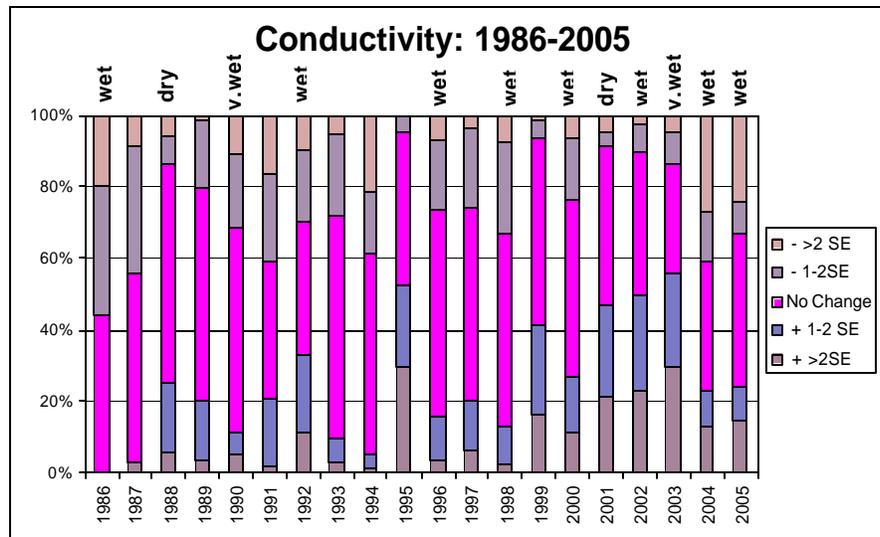


Figure 10a. Annual Change from "Normal" Conductivity in CSLAP Lakes (SE = Standard Error)

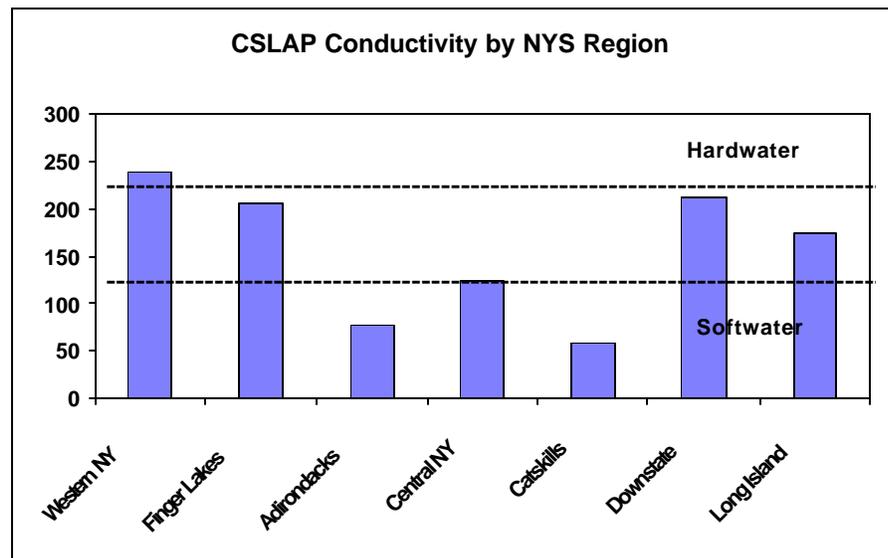


Figure 10b. Conductivity in CSLAP Lakes by NYS Region

Statewide Variability:

Although “hardwater” and “softwater” is not consistently defined by conductivity, in general lakes in the Adirondacks and Catskills have lower conductivity (softer water), and lakes downstate, in Western NY, and in the Finger Lakes region have higher conductivity (harder water). These regional differences are due primary to surficial geology and “natural” conditions in these areas. However, within each of these broad geographical areas, there are usually some lakes with higher conductivity and some lakes with lower conductivity readings.

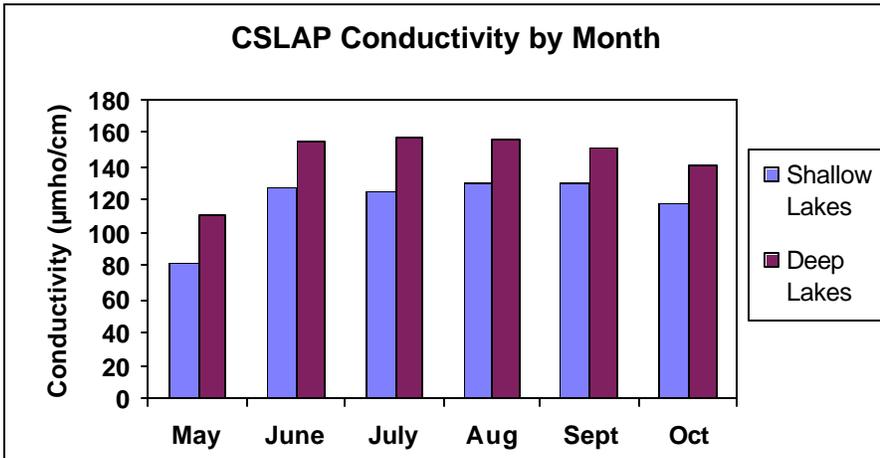


Figure 10c. Conductivity in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

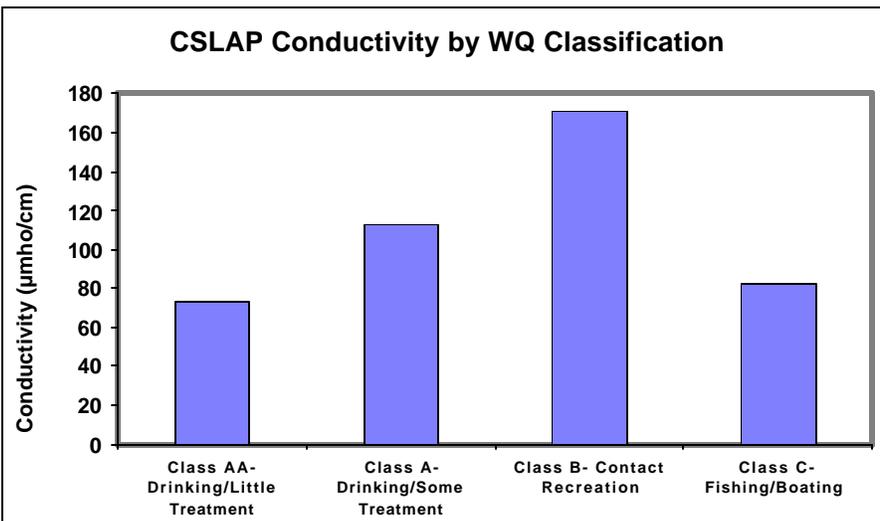


Figure 10d. Conductivity in CSLAP Lakes by Lake Use

Conductivity readings are substantially higher for lakes used primarily for contact recreation (Class B), and somewhat higher for lakes used for drinking water with some treatment (Class A). However, this is probably more reflective of geographical differences (there are relatively more softwater CSLAP lakes in the Adirondacks, which tend to have more Class A or Class AA lakes, at least in CSLAP, and more Class B lakes are found in hardwater regions) than any *de facto* connection between conductivity and lake usage.

Seasonal Variability:

Conductivity readings are much higher in the summer than in the late spring in many CSLAP lakes. These readings decreased in deep lakes in the summer and fall, but remained fairly steady in shallow lakes over this period (actual readings within specific lakes, however, may often vary significantly from week to week). Although lake destratification (turnover) brings bottom waters with higher conductivity to the lake surface in deeper lakes, conductivity readings dropped in the fall. It is possible that fully mixed conditions may be missed in some NYS lakes by discontinuing sampling after the end of October. Conductivity readings overall were higher in deep lakes, although this is may be an artifact of the sampling set (there are more CSLAP deep lakes in areas that “naturally” have harder water)

Lake Use Variability

Color

Annual Variability

The color of most CSLAP lakes has varied from year to year. The years with the lowest color readings, 1993 and 1995, had “normal” levels of precipitation, although four of the years with the highest color readings (1992, and 2002 through 2005) were wet, and the least colored waters generally occurred during dry conditions. Most lake samples (92%) correspond to water color readings too low (< 30 ptu) to significantly influence water clarity. Color readings were much higher in 2004 than in any other CSLAP sampling season. Given that color readings were also higher in 2002 and 2003, the increase in color may be attributable in part to the shift in laboratories, which occurred prior to the 2003 sampling season.

What Was Expected in 2006?

As noted above, color readings have generally been higher during wet years, and readings have been higher in the last three years, perhaps due to slightly different analytical methodology. Since 2006 generally corresponded to a wet year, it is likely that color readings in 2006 will at least be higher than the long-term average, although readings may not be higher than in 2004, which was also generally a wet year.

And What Happened at Lake Oscaleta in 2006?

2006 was the first year of CSLAP sampling at Lake Oscaleta. Color readings in 2006 were comparable to those measured in 1987 through the ALSC study.

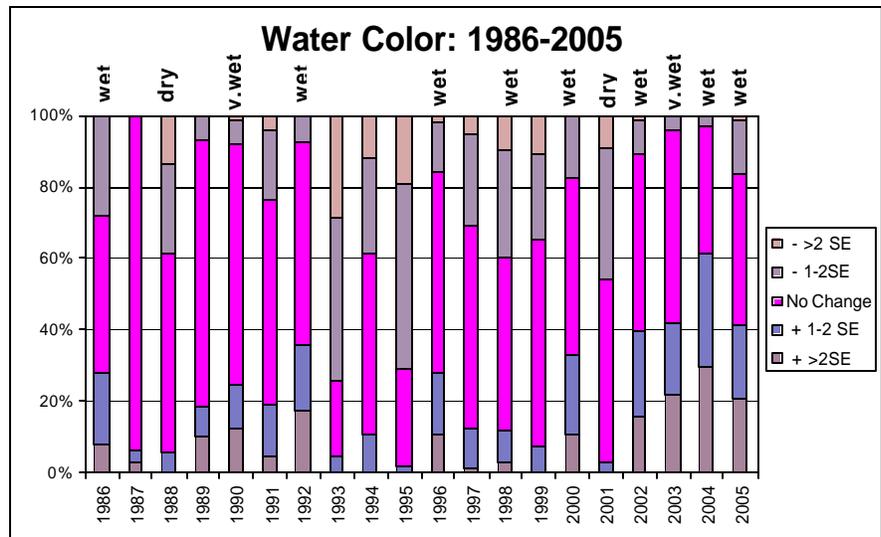


Figure 11a. Annual Change from “Normal” Color in CSLAP Lakes (SE = Standard Error)

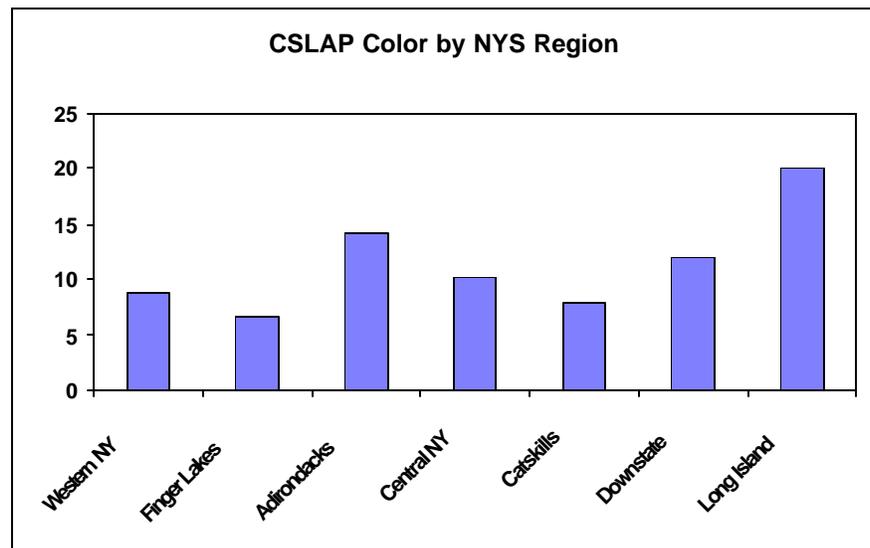


Figure 11b. Color in CSLAP Lakes by NYS Region

Statewide Variability:

Water color is highest in Long Island and the Adirondacks, and lowest in the Finger Lakes, Catskill and Western NY regions. This is mostly coincident with the statewide conductivity distribution (with softwater lakes more likely to be colored), and both seem to be largely consistent with the distribution of these lakes within New York State (in other words, the CSLAP dataset may be a representative cross-section of NYS lakes as related to color).

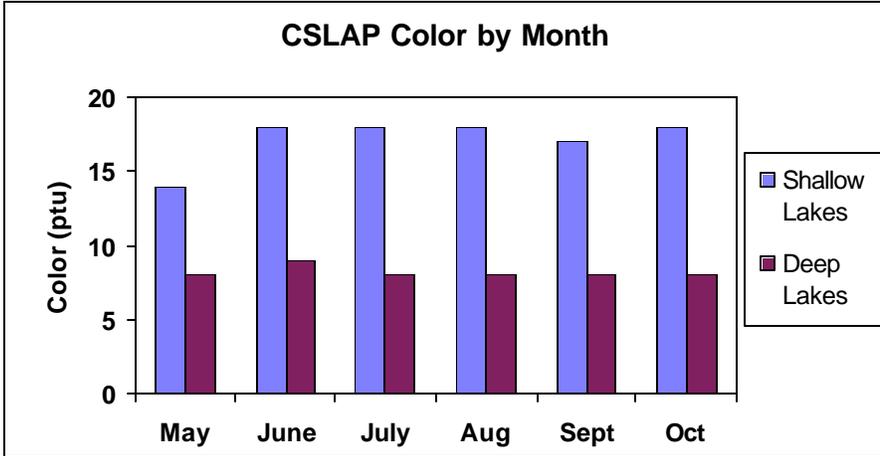


Figure 11c. Color in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

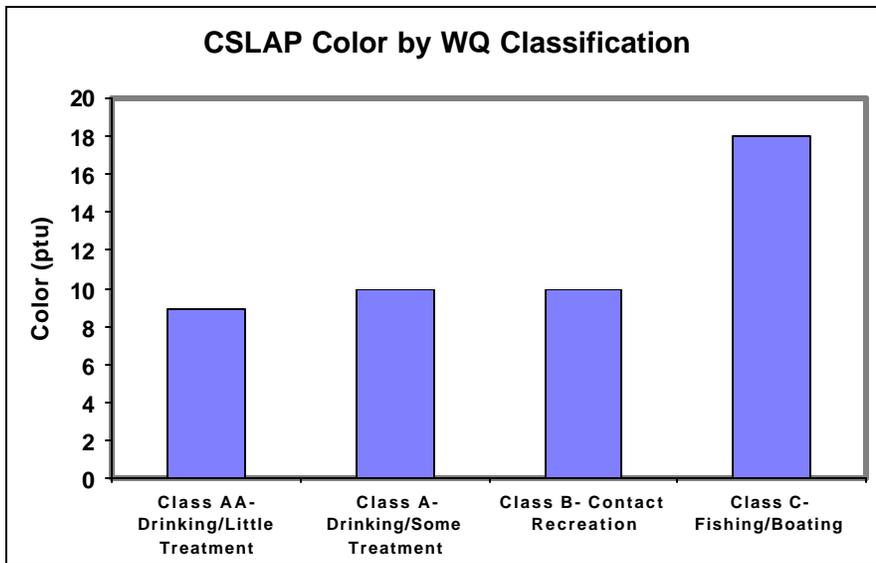


Figure 11d. Color in CSLAP Lakes by Lake Use

lakes (mean depth = 4 meters), while the other classes tend to be deeper lakes (mean depth = 9 meters). However, the elevated color readings correspond to elevated levels of dissolved organic matter, and may also reflect impediments (via economically viable water treatment, aesthetics, and potential formation of hazardous compounds during chlorination) to the use of these waters for potable water.

Seasonal Variability:

Color readings are significantly higher in shallow lakes than in deepwater lakes; these readings increase from spring to summer in these shallower lakes (perhaps due to dissolution of organic material, including algae, and wind-induced mixing during the summer) and then drop off again in late summer into the fall. Color generally follows the opposite trend in deeper lakes, with slightly decreasing levels perhaps due to more particle setting in the summer and remixing in the fall, although the seasonal trend in the deeper lakes is not as pronounced as in shallow lakes.

Lake Use Variability

Color readings are substantially higher for lakes used primarily for non-contact recreation (Class C), but this is probably more reflective of morphometric differences, for Class C lakes tend to be shallow

Nitrate

Annual Variability

Evaluating nitrate in CSLAP lakes is confounded by the relative lack of nitrate data for many sampling seasons (it was analyzed in water samples at a lower frequency, or not at all, in many years), the high number of undetectable nitrate readings, and some changes in detection levels. The limited data indicated that nitrate was highest in 1986 and 1989, two early CSLAP years in which nitrate was analyzed more frequently (including a relatively large number of early season samples), and in 2004 and 2005, which corresponded to the use of a new analytical tool. Readings were lowest in 1995, 2002 and 2003. Although nitrate levels are probably closely related to winter and spring precipitation levels (due to the higher nitrate readings in snowpacks), this is not apparent from Figure 12a. No readings have approached the state water quality standard (= 10 mg/l) in any CSLAP sample.

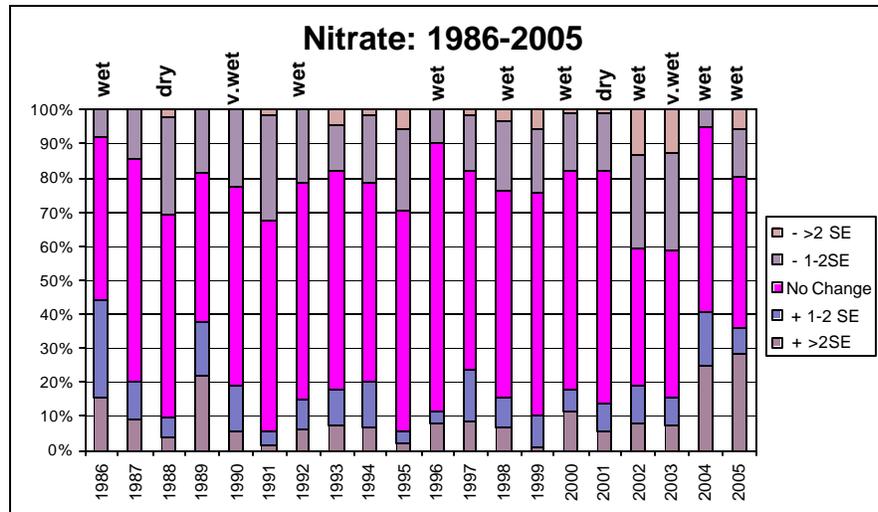


Figure 12a. Annual Change from "Normal" Nitrate in CSLAP Lakes (SE = Standard Error)

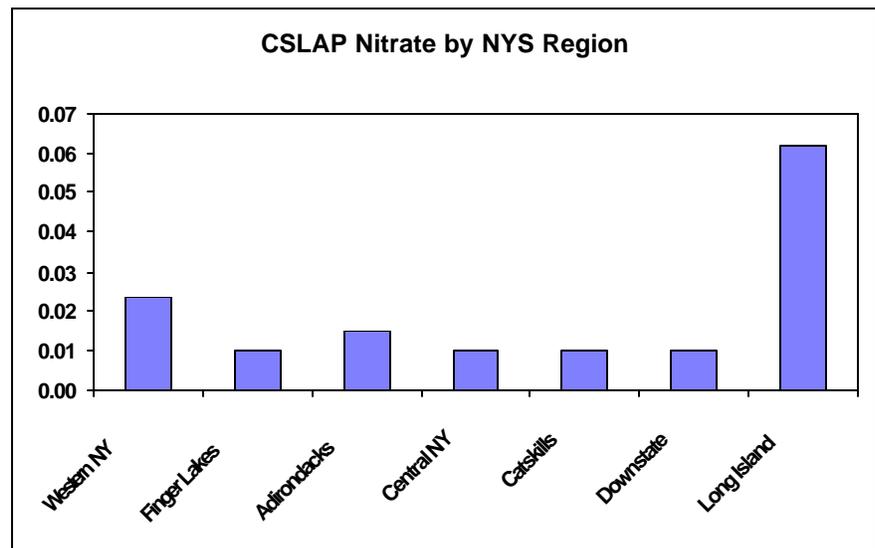


Figure 12b. Nitrate in CSLAP Lakes by NYS Region

What Was Expected in 2006?

Nitrate readings have been very unpredictable, although at nearly all times, all nitrate readings are small. Given the higher readings found in 2004, it is presumed that nitrate readings may also be slightly higher in 2006.

And What Happened at Lake Oscaleta in 2006?

2006 was the first year of CSLAP sampling at Lake Oscaleta. Nitrate readings in 2006 were probably comparable to those measured in 2003 (Cedar Eden) and 1987 (ALSC).

Statewide Variability:

Nitrate levels are highest in Long Island, Western NY, and the Adirondacks, and lowest in the other NYS regions. However, none of these regions demonstrate readings that are particularly high. Readings from individual lakes in the Long Island, Madison County, and the Adirondacks (spring only) are often elevated, although still well below water quality standards.

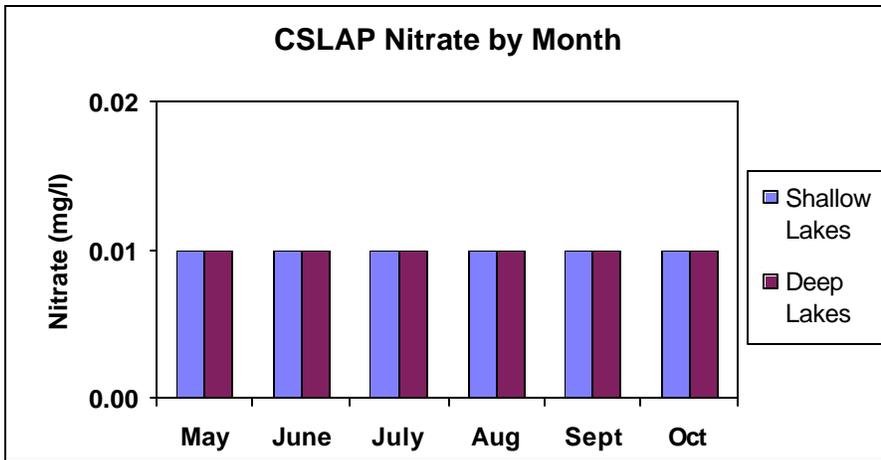


Figure 12c. Nitrate in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

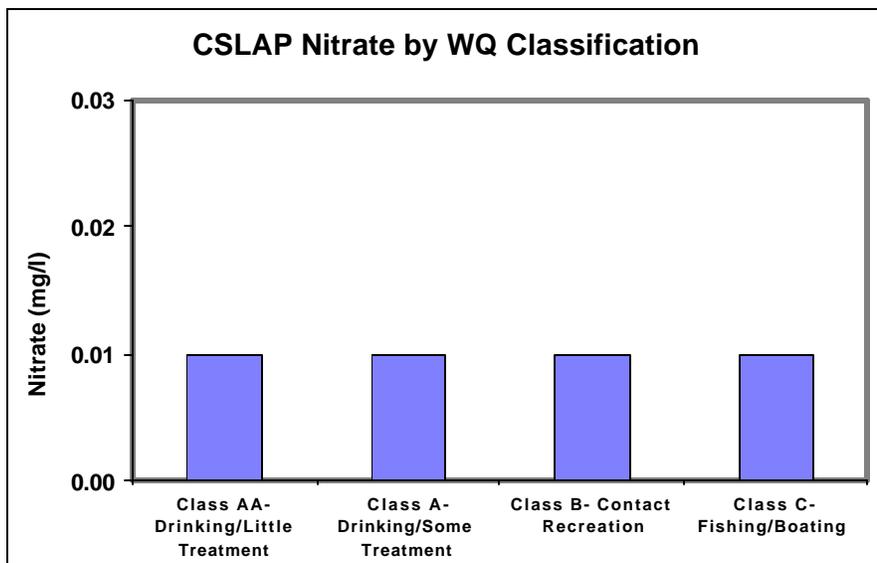


Figure 12d. Nitrate in CSLAP Lakes by Lake Use

undetectable nitrate readings).

Seasonal Variability:

Nitrate readings are not seasonally variable on a program-wide basis, as indicated in Figure 12c. However, in some individual lakes, in the regions listed above, nitrate is often detectable until early summer, and then undetectable through the rest of the sampling season (the large number of lakes with undetectable nitrate levels throughout the year overwhelm the statistics in Figure 12c).

Lake Use Variability

Nitrate readings appeared to be identical for all classes of lake uses, as indicated in Figure 12d. Higher early season nitrate readings are found in some lakes influenced by the melting of large winter snowpacks, such as some Class AA and A lakes in the Adirondacks, but these statistics cannot be easily teased from datasets strongly influenced by the large number of lakes with

**Trophic Indicators:
Water Clarity**

Annual Variability

Water clarity (transparency) has varied annually in most CSLAP lakes. There does not appear to be much of a correlation between clarity and precipitation- the highest clarity occurred in 1995, 1997, and 1999, which corresponded to normal precipitation (statewide), although the lowest clarity occurred during two wet years (1996 and 2000). There are no significant broad statewide water clarity trends, although (as described in other portions of this report), clear trends do exist on some lakes. The majority of water clarity readings in CSLAP lakes (56%) correspond to *mesotrophic* conditions (clarity between 2 and 5 meters), with 27% corresponding to *eutrophic* conditions ($Z_{sd} < 2$) and 17% corresponding to *oligotrophic* conditions ($Z_{sd} > 5$).

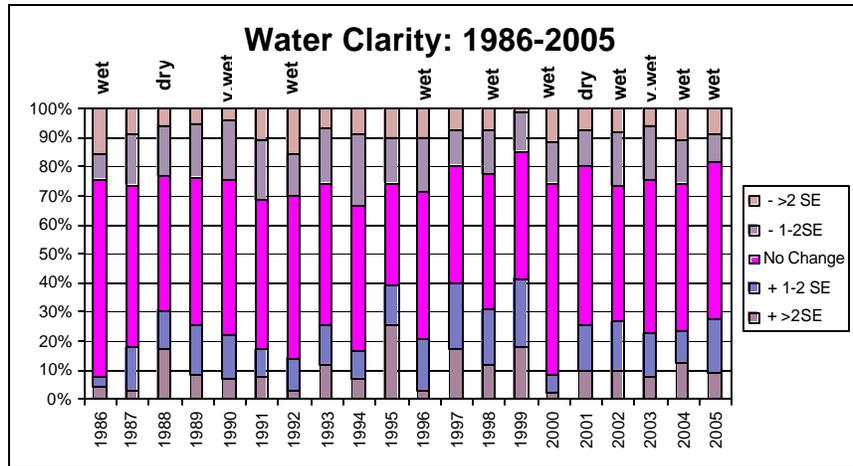


Figure 13a. Change from "Normal" Water Clarity in CSLAP Lakes (SE = Standard Error)

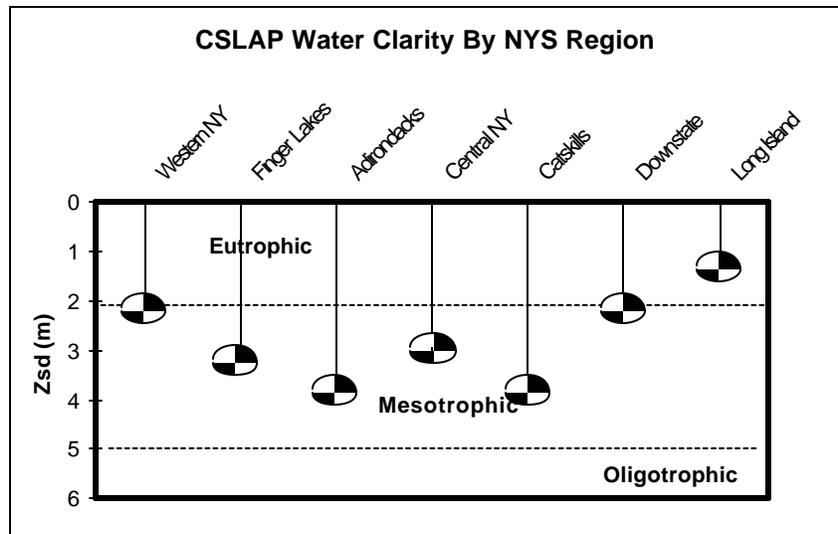


Figure 13b. Water Clarity in CSLAP Lakes by NYS Region

What Was Expected in 2006?

Since there is not a strong correlation between weather and water transparency readings, it is difficult to identify expected conditions. However, since water clarity seems to be lowest during wet years, it is likely that more lakes would exhibit slightly lower water transparency readings in 2006.

And What Happened at Lake Oscaleta in 2006?

2006 was the first year of CSLAP sampling at Lake Oscaleta. Water transparency readings were lower in 2006 than in either 2003 or 1987, but this is consistent with the higher nutrient and algae levels measured in 2006.

Statewide Variability:

As expected, water clarity is highest in the Adirondacks, Catskills, and Finger Lakes regions, and lowest in Long Island, Downstate, and Western NY. The differences are more pronounced (at least for the Adirondacks) when “naturally” colored lakes are not considered. However, except for Long Island (for which water clarity is at least partially limited by the shallow water depth), the “typical” lake in each of these regions would be classified as *mesotrophic*.

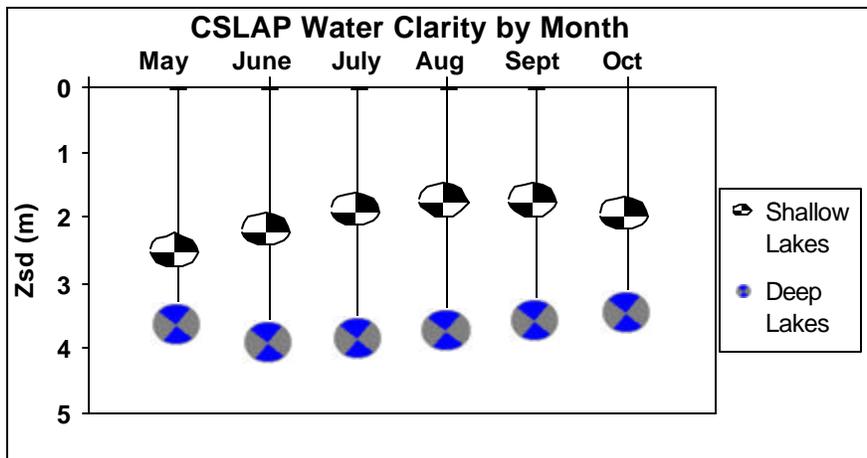


Figure 13c. Water Clarity in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

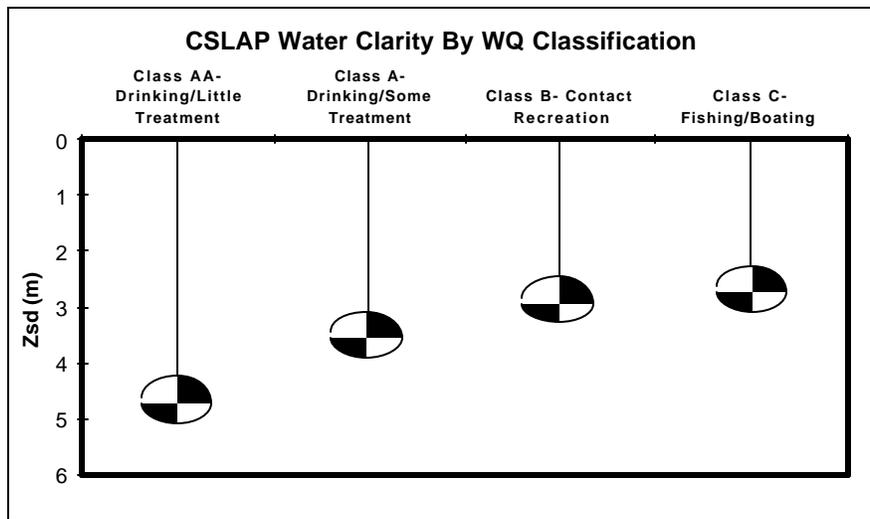


Figure 13d. Water Clarity in CSLAP Lakes by Lake Use

Seasonal Variability:

Water clarity readings are lower, as expected, in shallow lakes, even when water depth does not physically limit a water clarity measurement. Transparency decreases in both shallow and deep lakes over the course of the sampling season (the drop in clarity in shallower lakes is somewhat more significant), although clarity readings increase from spring to early summer in deeper CSLAP lakes. Water transparency rebounds slightly in shallower lakes in the fall, probably due to a drop in nutrient levels. The lack of “rebound” in deeper lakes may be due to occasional fall algal blooms in response to surface nutrient enrichment after lake turnover (see below)

Lake Use Variability

Water transparency decreases as the “sensitivity” of the lake use decreases, with higher clarity found in lakes used for potable water (Class

AA), and lower clarity found in lakes used primarily for contact and non-contact (fishing and boating) recreation. As with many of the other water quality indicators, this is due to both geographical and morphometric (depth) differences, although the original designation of these uses may also reflect these measurable and visually apparent water quality differences.

**Trophic Indicators:
Phosphorus (TP)**

Annual Variability

Total phosphorus (TP) has varied annually in most CSLAP lakes. The highest phosphorus readings occurred during 1991, 1996, 1998, 2000, and 2003, the latter four of which corresponded to wet years. However, the lowest readings, from 1989, 1997, and 2002, did not correspond to unusually dry years, and 2004 was a fairly wet year. The majority of phosphorus readings in CSLAP lakes (39%) correspond to *mesotrophic* conditions (clarity of 2 to 5m), with 27% corresponding to *eutrophic* conditions (< 2m clarity) and 34% corresponding to *oligotrophic* conditions (> 5m clarity); the latter is a much higher percentage than the trophic designation for water clarity.

What Was Expected in 2006?

As noted above, there is not a strong correlation between weather and total phosphorus, and there does not appear to be a consistent long-term pattern in the total phosphorus data.

The data also does not appear to be significantly laboratory-dependent, at least as apparent in Figure 14a. As such, it is difficult to predict whether phosphorus levels might be expected to be higher or lower in most CSLAP lakes in 2006.

And What Happened at Lake Oscaleta in 2006?

2006 was the first year of CSLAP sampling at Lake Oscaleta. Phosphorus readings in 2006 were higher than those measured in either 2003 (Cedar Eden) or 1987 (ALSC), although readings in all three years were typical of highly productive lakes.

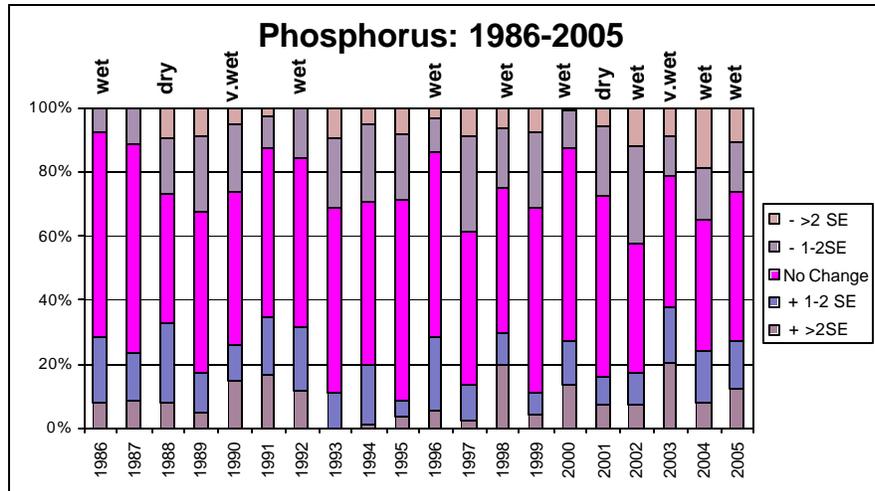


Figure 14a. Annual Change from "Normal" TP in CSLAP Lakes (SE = Standard Error)

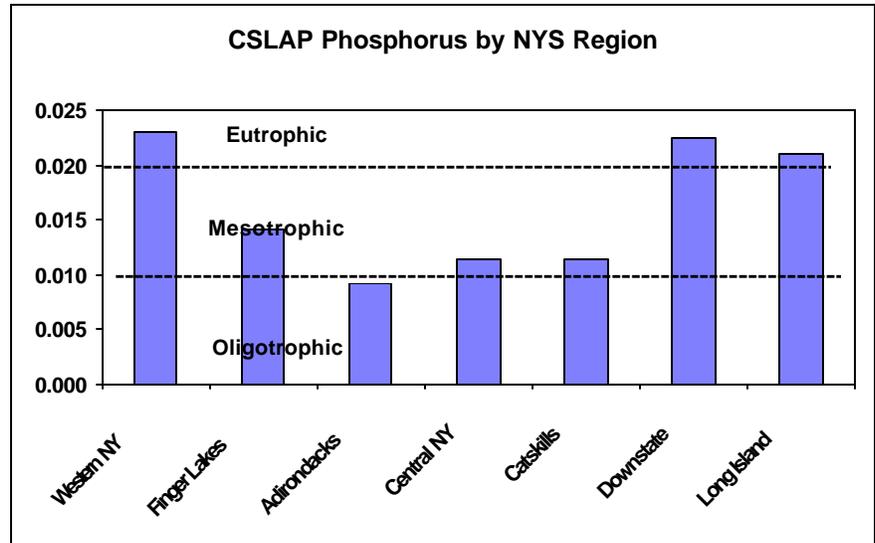


Figure 14b. TP in CSLAP Lakes by NYS Region

Statewide Variability:

As expected, nutrient levels are lowest in the Adirondacks, Catskills, and Central New York (where clarity is highest) and highest in Long Island, Downstate, and Western NY, where clarity is lowest. In the latter three regions, the “typical” lake in each of these regions would be classified as *eutrophic*, while only in the Adirondacks could most lakes be described as *oligotrophic*, based on nutrients.

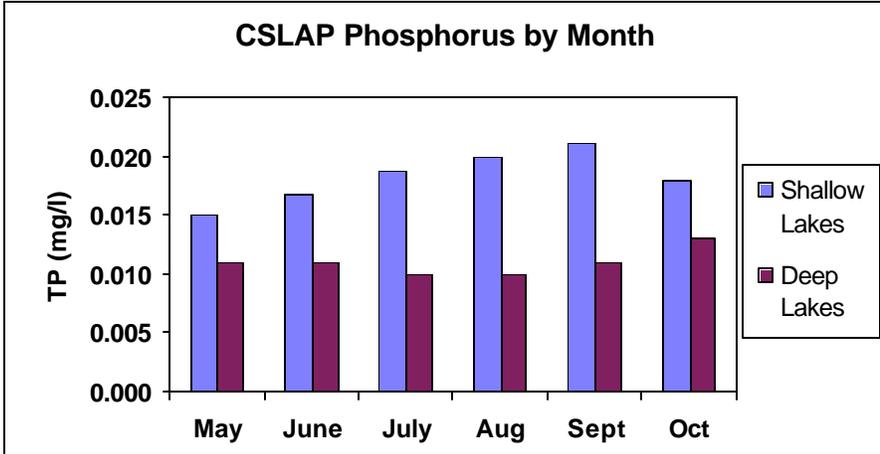


Figure 15c. TP in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

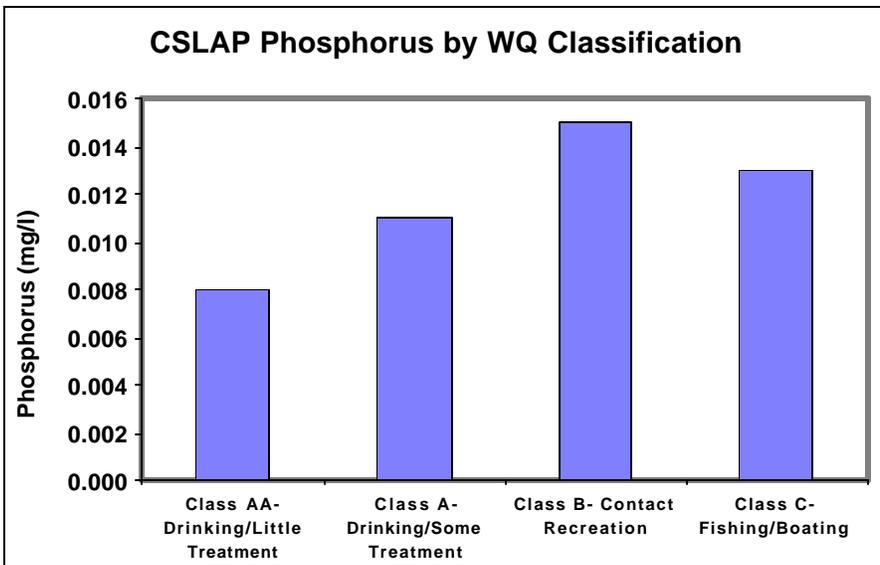


Figure 15d. TP in CSLAP Lakes by Lake Use

than Class C lakes (contact recreation versus non-contact recreation), these lakes actually have higher nutrient levels, perhaps reflecting the influence of deepwater nutrient enrichments (these lakes are typically deeper) and the “unofficial” use of Class C waters for bathing and contact recreation.

Seasonal Variability:

Nutrient levels are higher, as expected, in shallow lakes, and phosphorus levels increase in shallow lakes over the course of the sampling season, until dropping in the fall. However, phosphorus levels in deeper lakes are lower and decrease slightly through July, then increase into the fall. The latter phenomenon is due to surface nutrient enrichment after lake turnover (high nutrient water from the lake bottom, due to release of nutrients from poorly oxygenated lake sediments in the summer, migrates to the lake surface when the lake destratifies).

Lake Use Variability

Phosphorus readings are lower in lakes used for minimally treated potable water intakes (Class AA), and are higher for other lake uses. Although Class B waters are utilized for a “higher” lake use

Trophic Indicators: Chlorophyll *a* (Chl.a)

Annual Variability

Chlorophyll *a* (Chl.a) has varied in most CSLAP lakes more significantly than the other trophic indicators, as is typical of biological indicators (which tend to grow “patchy”). With the exception of the very high readings in 1987 (probably due to a lab “problem”), the highest chlorophyll *a* levels occurred during 1990, 1991, 1996, and 2000, corresponded to wet years. However, the lowest readings, from 1989, 1997, and 2001 through 2005 also corresponded to normal to wet conditions in most of these years. The consistently lower chlorophyll readings in the last four years may also correspond to the shift in laboratories, although both labs use the same analytical methodology. The near majority of chlorophyll readings in CSLAP lakes (49%) correspond to *mesotrophic* conditions (clarity between 2 and 5 meters), with 33% corresponding to *eutrophic* conditions ($Z_{sd} < 2$) and 18% corresponding to *oligotrophic* conditions ($Z_{sd} > 5$); these percentages are more like those for water clarity rather than those for phosphorus.

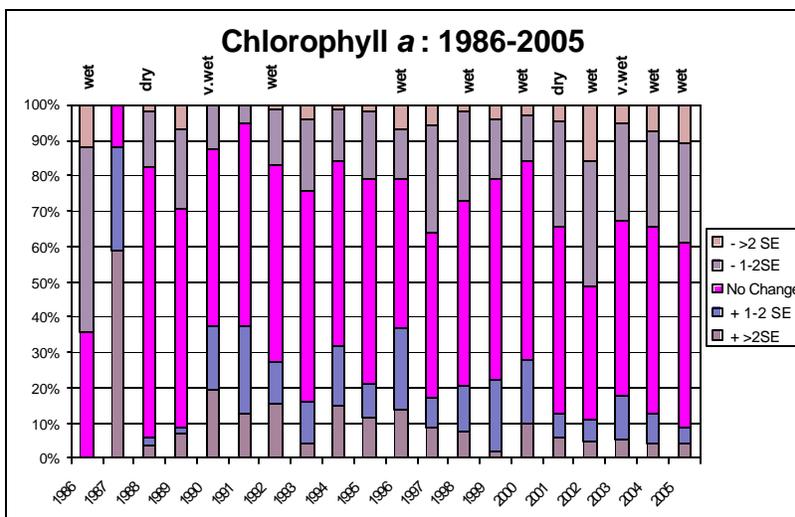


Figure 16a. Annual Change from “Normal” Chlorophyll *a* in CSLAP Lakes (SE = Standard Error)

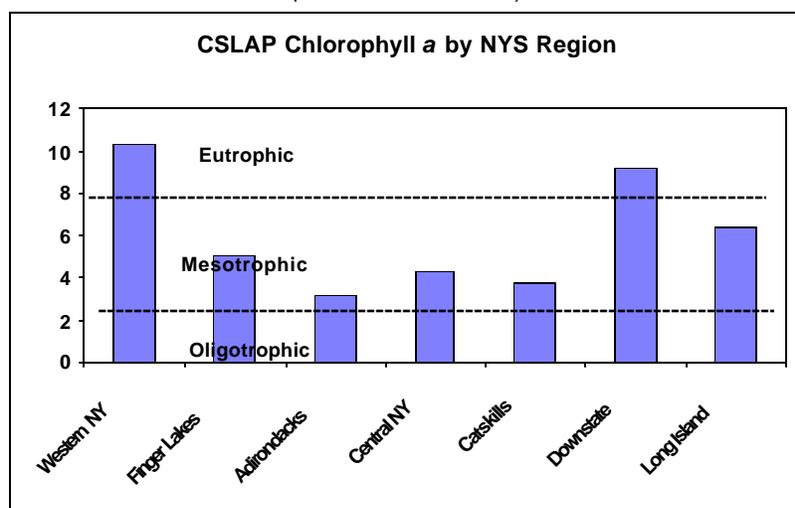


Figure 16b. Chlorophyll *a* in CSLAP Lakes by NYS Region

What Was Expected in 2006?

It is likely that chlorophyll readings would be lower than the long-term average for most CSLAP lakes in 2006, due to consistently lower readings coming from the same laboratory in the last several years. Since 2006 was also generally a hotter and “stickier” year than is typical at most NYS lakes, it would not be surprising to see higher-than-usual chlorophyll readings, at least relative to the last several years. However, this did not consistently occur in previously hot/humid years.

And What Happened at Lake Oscaleta in 2006?

2006 was the first year of CSLAP sampling at Lake Oscaleta. Chlorophyll *a* readings were higher in 2006 than in 2003, the only other year in which these data were collected. This is consistent with the higher phosphorus readings and lower water transparency in 2006.

Statewide Variability:

As with phosphorus, chlorophyll levels are lowest in the Adirondacks, Central New York, and the Catskills (where clarity is highest) and highest in Long Island, Downstate, and Western NY, where clarity is lowest. In the latter two regions, the “typical” lake in each of these regions would be classified as *eutrophic*, while lakes in the other regions would be described as *mesotrophic*.

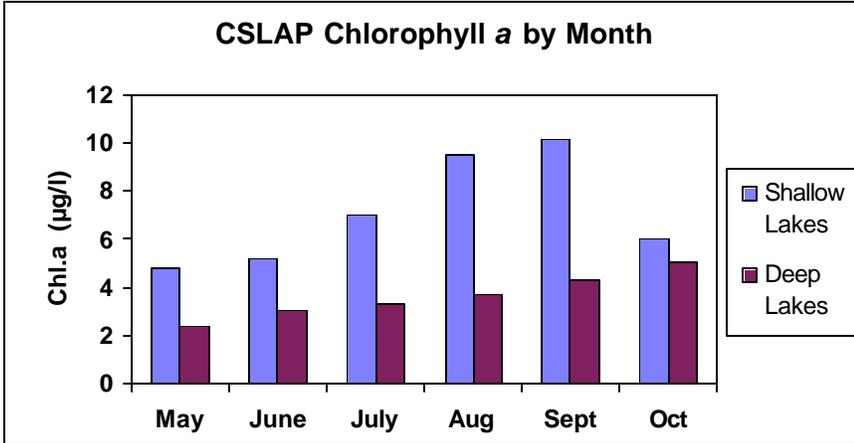


Figure 16c. Chlorophyll a in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

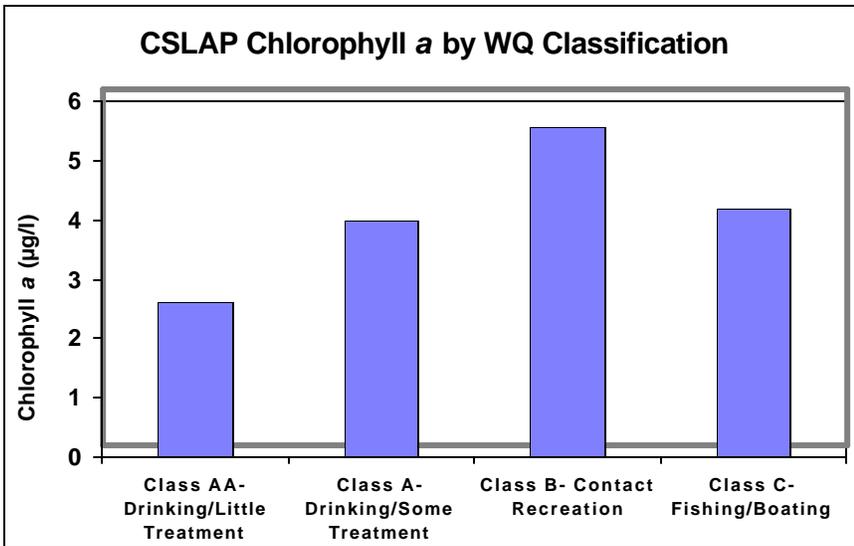


Figure 16d. Chlorophyll a in CSLAP Lakes by Lake Use

Seasonal Variability:

Chlorophyll levels are higher, as expected, in shallow lakes, and increase in both shallow and deep lakes over the course of the sampling season, with chlorophyll readings dropping in shallow lakes in the fall. The steady increase in chlorophyll in both shallow and (to a lesser extent) deep lakes is consistent with the change in phosphorus over the same period, due to steady migration of nutrients released from poorly oxygenated lake sediments during the summer and especially in the fall (as well as drier weather, increased lake use, and other factors).

Lake Use Variability

Chlorophyll readings are lower in lakes used for minimally treated potable water intakes (Class AA), and are higher for other lake uses. Although Class B waters are utilized for a “higher” lake use than Class C lakes (contact recreation versus non-contact recreation), these lakes

actually have similar levels, perhaps reflecting the influence of deepwater nutrient enrichments (these lakes are typically deeper) and the “unofficial” use of Class C waters for bathing and contact recreation. This is similar to the use pattern for phosphorus.

Water Quality Assessment (QA on the Perception Form)

Annual Variability

Water quality assessments (the perceived physical condition of the lake, or QA on the use impairment surveys) were least favorable in the very wet (2000) and very dry (1995) years, suggesting the lack of correlation between weather and perceived water quality (although 1995 was also the year with the most “improved” conditions). The general perception of CSLAP lakes in 2005 indicated no strong changes in perceived water quality, whether favorable or unfavorable. Although there is a strong connection between measured and perceived water clarity in most CSLAP lakes, this is not closely reflected in the comparison of Figures 14a and 17a.

What Was Expected in 2006?

There was not a strong connection between precipitation (within mostly normal weather patterns) and perceived water quality, or even between measured and perceived water quality conditions. As such, it is difficult to identify expected conditions in 2006, although since water clarity readings were mostly within normal ranges, it is reasonable to expect that perceived water quality conditions would also largely be unchanged.

And What Happened at Lake Ooscaleta in 2006?

2006 was the first year of CSLAP sampling at Lake Ooscaleta. Given the lack of historical data for the lake (or the lack of data available to the NYSDEC), a comparison of 2006 data to those collected in previous years is not possible.

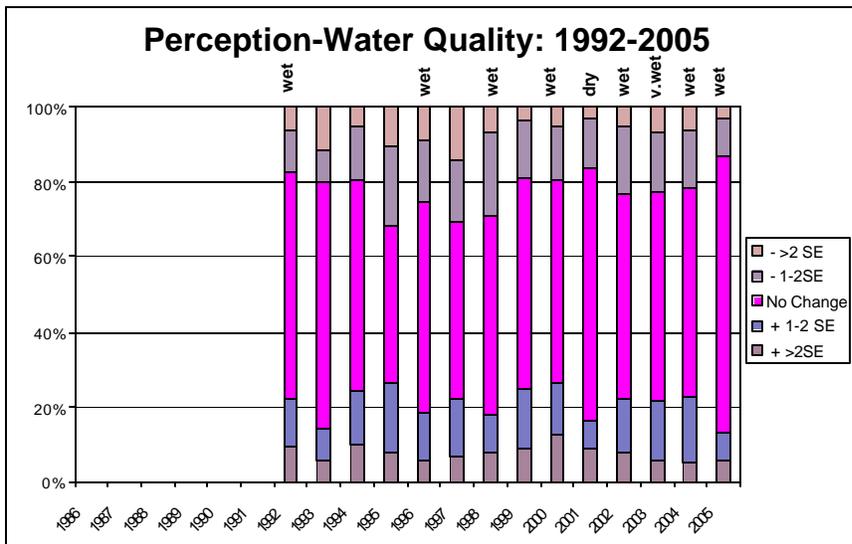


Figure 17a. Annual Change from “Normal” Water Quality Assessment in CSLAP Lakes (SE = Standard Error)

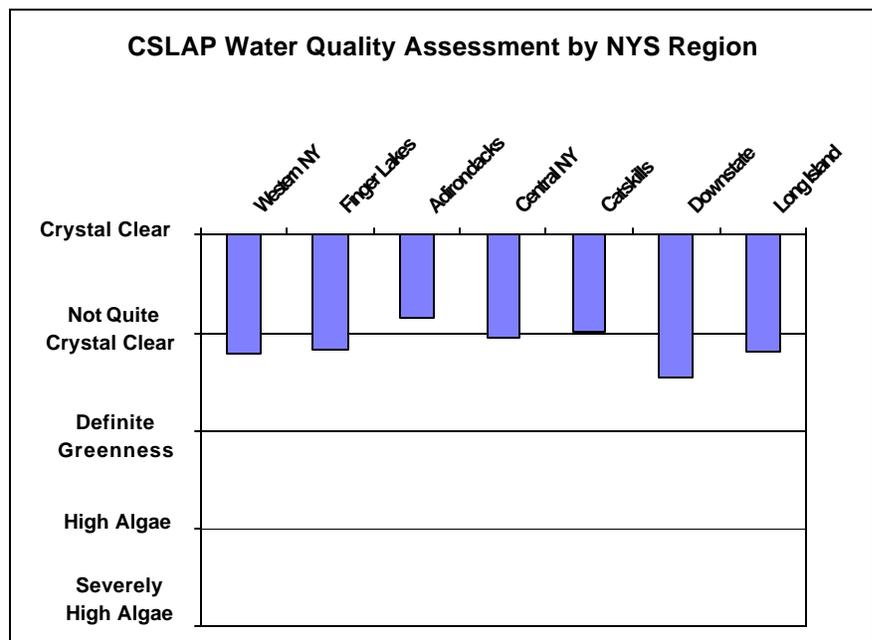


Figure 17b. Water Quality Assessment in CSLAP Lakes by NYS Region

Statewide Variability:

The most favorable water quality assessments (at least in support of contact recreation) occurred in the Adirondacks, Catskills, and Central New York, as expected, and water quality assessments were slightly less favorable in Downstate, Western NY, and Long Island. This is mostly consistent with the water clarity readings in these regions. However, since the difference between the most favorable (Adirondacks) and least favorable (Downstate) assessments is smaller than the measured water transparency differences, this suggests that the relatively low water clarity in the latter regions may be considered “normal” by lake residents.

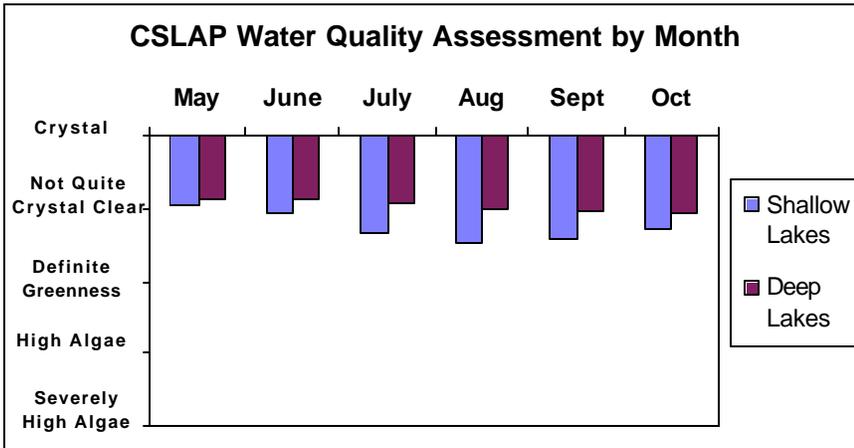


Figure 17c. Water Quality Assessment in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

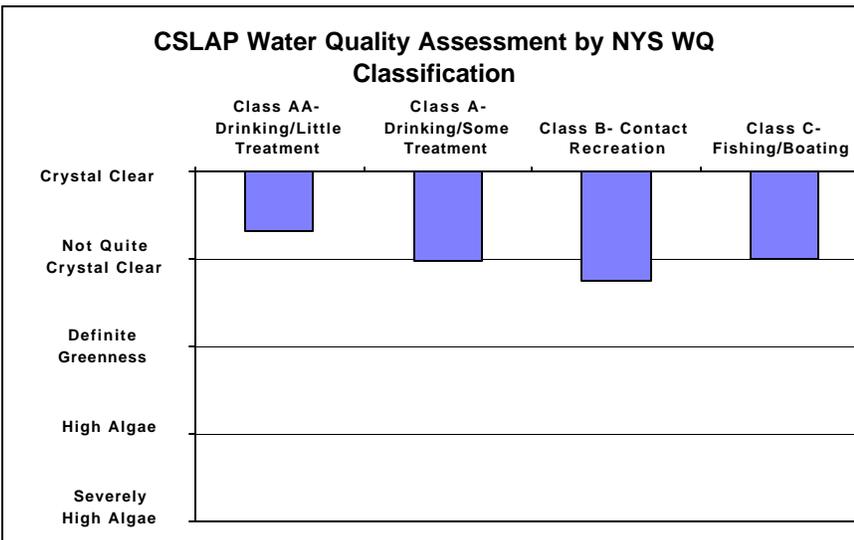


Figure 17d. Water Quality Assessment in CSLAP Lakes by Lake Use

assessments, perhaps reflecting the influence of deepwater nutrient enrichments (these lakes are typically deeper) and the “unofficial” use of Class C waters for bathing and contact recreation. This is similar to the pattern seen for the trophic indicators.

Seasonal Variability:

Water quality assessments become less favorable as the summer progresses in both deep and (especially) shallow lakes, coincident with similar patterns for the trophic indicators. However, the seasonal changes in these assessments are not very large. These assessments become slightly more favorable in shallow lakes in the fall, consistent with the improved (measured) water clarity, although overall water quality assessments are less favorable all year in shallow lakes.

Lake Use Variability

Water quality assessments are more favorable in lakes used for potable water intakes (Class AA and Class A), and less favorable for other lake uses. Although Class B waters are utilized for a “higher” lake use than Class C lakes (contact recreation versus non-contact recreation), these lakes actually have similar water quality

Aquatic Plant (Weed) Assessment (QB)

Annual Variability

Aquatic plant assessments (the perceived extent of weed growth in the lake, or QB on the use impairment surveys) indicated that weeds grew most significantly in 1995 (normal conditions) and 2000 (wet conditions), and weed growth was less extensive in 1994 and 1999, suggesting the lack of correlation between weather and weed densities. The highest weed growth occurred when the perceived physical condition (clarity) of the lake was also least favorable- these conditions may offer a selective advantage to invasive or exotic weeds (such as *Myriophyllum spicatum*).

What Was Expected in 2006?

There was not a strong connection between precipitation and extent of weed growth, at least as measurable through CSLAP. This makes it difficult to identify expected conditions in 2006. As is always the case, it is likely that the extent of weed growth in any particular CSLAP lake in 2006 is unrelated to the extent of weed growth in most other NYS lakes, and is not readily predictable given historical patterns of aquatic plant growth in that lake.

And What Happened at Lake Ooscaleta in 2006?

2006 was the first year of CSLAP sampling at Lake Ooscaleta. Given the lack of historical data for the lake (or the lack of data available to the NYSDEC), a comparison of 2006 data to those collected in previous years is not possible.

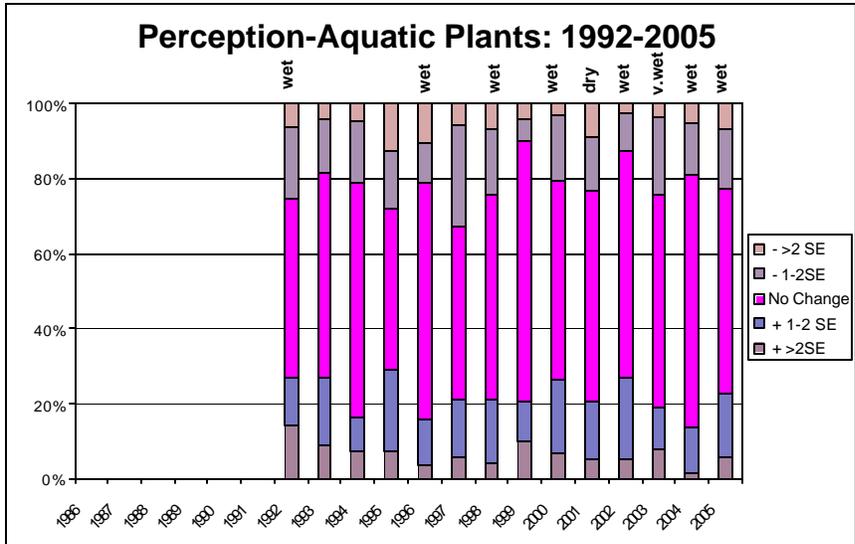


Figure 18a. Annual Change from "Normal" Weed Assessment in CSLAP Lakes (SE = Standard Error)

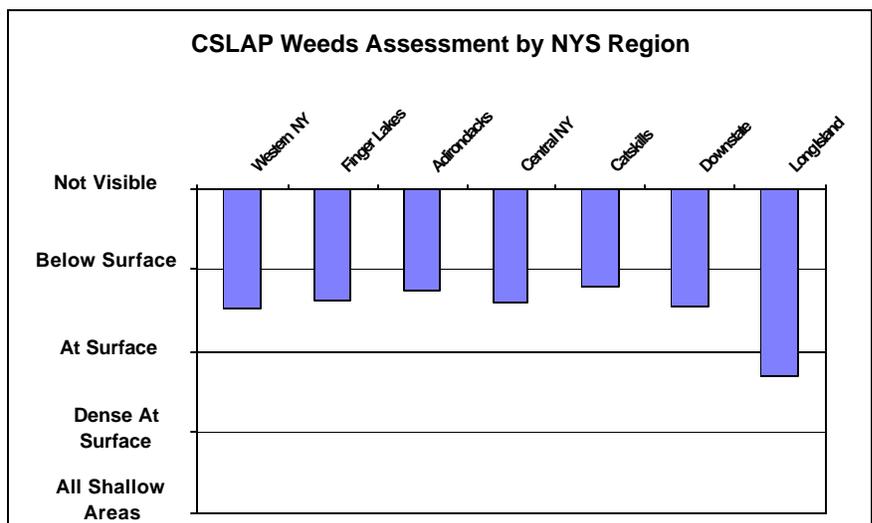


Figure 18d. Weed Assessment in CSLAP Lakes by NYS Region

Statewide Variability:

Aquatic plant growth was most significant in Long Island (and to a lesser extent Downstate and Western NY) and least significant in the Catskills and Adirondacks area. The former may have a larger concentration of shallow lakes (Long Island) or preponderance of exotic weeds (Downstate and Western NY), while the latter may correspond to deeper lakes or fewer instances of these invasive weeds, although it is also likely that invasive weed growth may be increasing in many lakes within these “less impacted” areas.

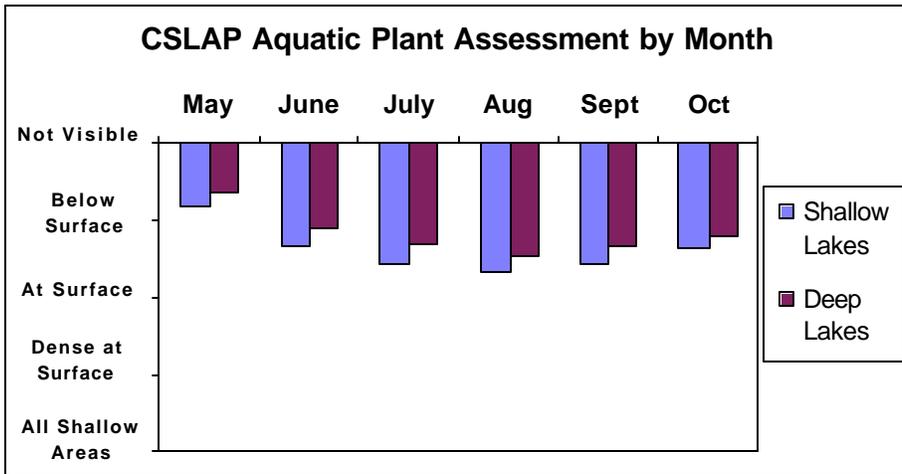


Figure 18c. Weed Assessment in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

Seasonal Variability:

As expected, aquatic plant densities and coverage increase seasonally (through late summer) in both shallow and deep lakes, with greater aquatic plant coverage and densities found in shallow lakes. Peak aquatic plant densities tend to occur in late summer. The variability from one lake to another (from very little growth to dense growth at the lake surface) is more pronounced later in the summer. Despite higher clarity in shallow lakes in the fall, aquatic plant coverage decreases, while the drop in fall plant coverage in deeper lakes is less pronounced.

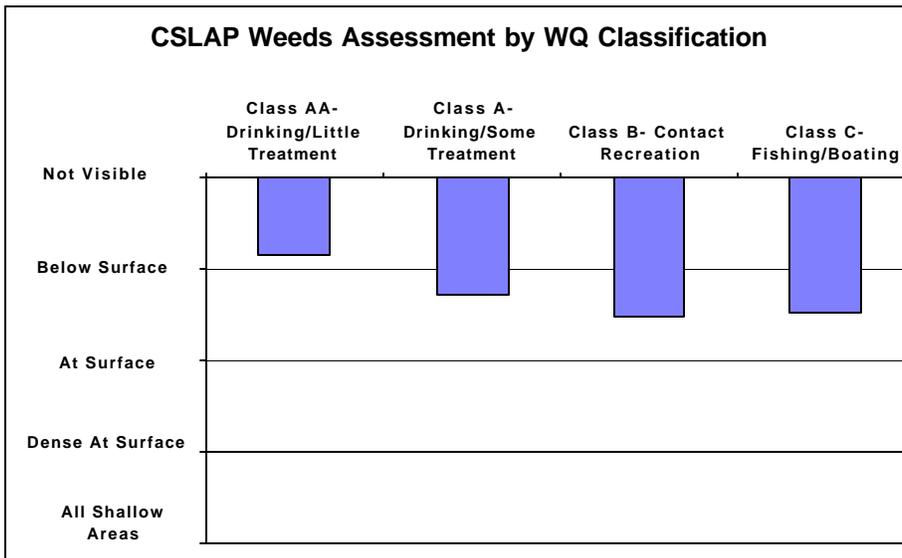


Figure 18d. Weed Assessment in CSLAP Lakes by Lake Use elevation areas in the Catskills and Adirondacks, and Class C lakes tend to be shallower than Class AA or Class A lakes).

Lake Use Variability

Aquatic plant coverage was more significant in Class B and Class C lakes than in other lakes, but this (again) is probably a greater reflection of geography or lake size and depth (Class B lakes tend to be found outside the high

Recreational Assessment (QC)

Annual Variability

Recreational assessments (the perceived recreational suitability of the lake, or QC on the use impairment surveys) have varied from year to year, with no clear long-term pattern. The most favorable assessments were in 1997, corresponding to the year with the lowest aquatic plant (weed) coverage. This was also among the years with the most favorable water quality assessments. The years with the most favorable water quality assessments (1995 and 1998) were among the years with the most favorable recreational assessments, despite higher than usual weed densities. This suggests that recreational assessments are influenced by both water quality conditions and aquatic plant densities. The extent of “normal” conditions (the middle bar in Figure 19a) has generally not changed significantly since perception surveys were first conducted in 1992.

What Was Expected in 2006?

There was not a strong connection between precipitation (within mostly normal weather patterns) and perceived water quality, or even between measured and perceived water quality conditions. As such, it is difficult to identify expected conditions in 2006, although since water clarity readings were mostly within normal ranges, it is reasonable to expect that perceived water quality conditions would also largely be unchanged.

And What Happened at Lake Ooscaleta in 2006?

2006 was the first year of CSLAP sampling at Lake Ooscaleta. Given the lack of historical data for the lake (or the lack of data available to the NYSDEC), a comparison of 2006 data to those collected in previous years is not possible.

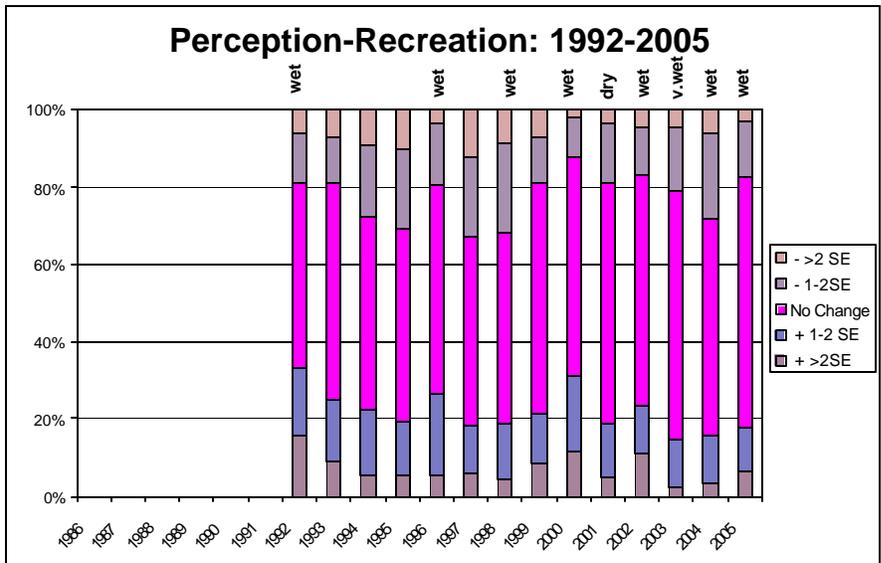


Figure 19a. Annual Change from “Normal” Recreational Assessment in CSLAP Lakes (SE = Standard Error)

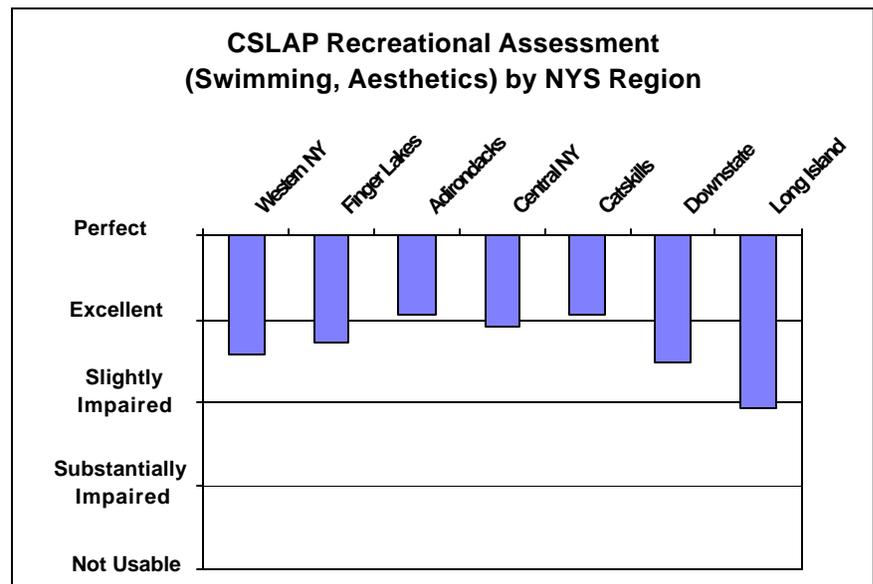


Figure 19b. Recreational Assessment in CSLAP Lakes by NYS Region

Statewide Variability:

Recreational assessments are most favorable in the Adirondacks and Catskills, and less favorable in Long Island and (to a lesser extent) Downstate and in Western New York. This appears to be in response to less favorable assessments of water quality and aquatic plant growth, respectively. Except for (the assessments in the small number of CSLAP lakes in) Long Island, overall recreational assessments in all regions are, in general, highly favorable.

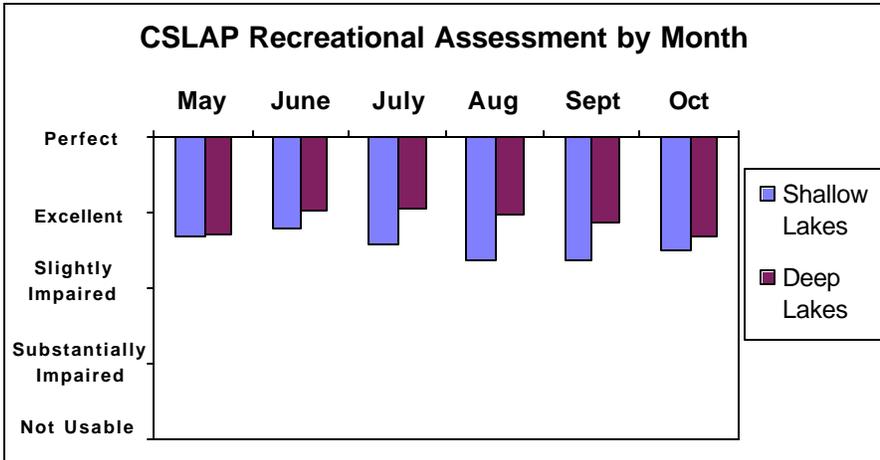


Figure 19c. Recreational Assessment in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

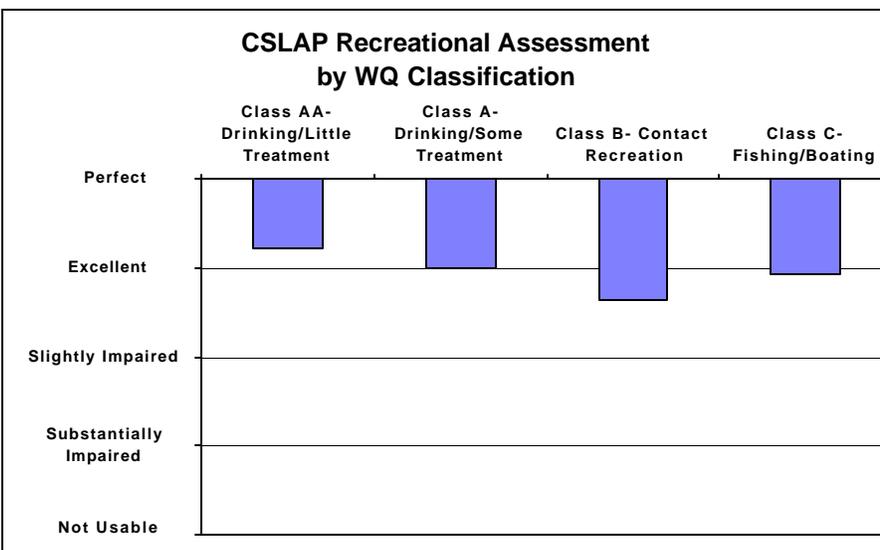


Figure 19d. Recreational Assessment in CSLAP Lakes by Lake Use less favorable than in Class AA and A lakes. This may be considered a validation of these classifications (recognizing, again, that many Class C lakes continue to fully support contact recreation and perhaps even potable water use).

Seasonal Variability:

Recreational assessment in both shallow and deep lakes tends to improve from spring to early summer, and then degrade through the summer, improving in shallow lakes in the fall. As expected, this generally corresponds to seasonal increases in aquatic plant coverage in deep lakes, and also to seasonally degrading water quality in shallow lakes. Overall recreational assessments are more favorable in deep lakes every month of the sampling season, although the differences are less pronounced in late spring and early fall (and winter, when every lake looks nice!)

Lake Use Variability

Recreational assessment becomes less favorable as the designated lake use becomes less sensitive (drinking water to contact recreation), although recreational assessments of Class C lakes are only slightly

VI. DETAILED LAKE OSCALETA WATER QUALITY DATA SUMMARY

CSLAP is intended to provide a database to help lake associations understand lake conditions and foster sound lake protection and pollution prevention decisions. This individual lake summary for **2006** contains two forms of information. The **raw data** and **graphs** present a snapshot or glimpse of water quality conditions at each lake. They are based on (at most) eight or nine sampling events during the summer. As lakes are sampled through CSLAP for a number of years, the database for each lake will expand, and assessments of lake conditions and water quality data become more accurate. For this reason, lakes new to CSLAP for only one year will not have information about annual trends.

Raw Data

Two “**data sets**” are provided below. The data presented in Table 1 include an annual summary of the minimum, maximum, and average for each of the CSLAP sampling parameters, including data from other sources for which sufficient quality assurance/quality control documentation is available for assessing the validity of the results. This data may be useful for comparing a particular data point for the current sampling year with historical data or information. Table 2 includes more detailed summaries of the 2006 and historical data sets, including some evaluation of water quality trends, comparison against existing water quality standards, and whether 2006 represented a typical year.

Graphs

The second form of data analysis for your lake is presented in the form of **graphs**. These graphs are based on the raw data sets to represent a snapshot of water quality conditions at your lake. The more sampling that has been done on a particular lake, the more information that can be presented on the graph, and the more information you have to identify annual trends for your lake. For example, a lake that has been doing CSLAP monitoring consistently for five years will have a graph depicting five years worth of data, whereas a lake that has been doing CSLAP sampling for only one year will only have one. Therefore, it is important to consider the number of sampling years of information in addition to where the data points fall on a graph when trying to draw conclusions about annual trends. There are certain factors not accounted for in this report that lake managers should consider:

- **Local weather conditions** (high or low temperatures, rainfall, droughts or hurricanes). Due to delays in receiving meteorological data from NOAA stations within NYS, weather data from individual weather stations or the present sampling season are not included in these reports. Some of the variability reported below can be attributed more to weather patterns than to a “real” water trend or change. However, it is presumed that much of the sampling “noise” associated with weather is dampened over multiple years of data collection, and thus should not significantly influence the limited trend analyses provided for CSLAP lakes with longer and larger databases.
- **Sampling season and parameter limitations**. Because sampling is generally confined to June-September, this report does not look at CSLAP parameters during the winter and other seasons. Winter conditions can impact the usability and water quality of a lake conditions. In addition, there are other sampling parameters (fecal coliform, dissolved oxygen, etc.) that may be responsible for chemical and biological processes and changes in physical measurements (such as water clarity) and the perceived conditions in the lake. **The 2006 CSLAP report attempts to standardize some comparisons by limiting the evaluation to the summer recreational season and the most common sampling periods (mid-June through mid-September), in the event that samples are collected at other times of the year (such as May or October) during only some sampling seasons.**

TABLE 1: CSLAP Data Summary for Lake Oseleta

Year	Min	Avg	Max	N	Parameter
2006-06	0.50	2.13	3.10	8	CSLAP Zsd
2006	0.50	2.13	3.10	8	CSLAP Zsd
2003	2.90	3.39	4.30	5	CedEd Zsd
1987	2.30	3.20	4.10	2	ALSC Zsd
Year	Min	Avg	Max	N	Parameter
2006-06	0.016	0.031	0.055	8	CSLAP Tot.P
2006	0.016	0.031	0.055	8	CSLAP Tot.P
2006	0.038	0.057	0.082	7	CSLAP HypoTP
2003	0.020	0.026	0.037	5	CedEd Tot.P
2003	0.050	0.071	0.085	5	CedEd HypoTP
1987	0.011	0.018	0.025	2	ALSC Tot.P
1987	0.027	0.027	0.027	1	ALSC HypoTP
Year	Min	Avg	Max	N	Parameter
2006-06	0.01	0.02	0.03	8	CSLAP NO3
2006	0.01	0.02	0.03	8	CSLAP NO3
2003		0.01			CedEd NO3
1987	0.01	0.01	0.01	2	ALSC NO3
Year	Min	Avg	Max	N	Parameter
2006-06	0.01	0.04	0.12	8	CSLAP NH4
2006	0.01	0.04	0.12	8	CSLAP NH4
1987	0.01	0.01	0.01	2	ALSC NH4
Year	Min	Avg	Max	N	Parameter
2006-06	0.48	0.57	0.80	5	CSLAP TDN
2006	0.48	0.57	0.80	5	CSLAP TDN
Year	Min	Avg	Max	N	Parameter
2006-06	10	20	30	5	CSLAP TN/TP
2006	10	20	30	5	CSLAP TN/TP
Year	Min	Avg	Max	N	Parameter
2006-06	12	20	35	8	CSLAP TColor
2006	12	20	35	8	CSLAP TColor
1987	15	18	20	2	ALSC TColor
Year	Min	Avg	Max	N	Parameter
2006-06	7.39	8.32	9.36	8	CSLAP pH
2006	7.39	8.32	9.36	8	CSLAP pH
2003	7.40	7.66	8.00	5	CedEd pH
1987	7.59	7.78	7.97	2	ALSC pH
Year	Min	Avg	Max	N	Parameter
2006-06	121	133	151	8	CSLAP Cond25
2006	121	133	151	8	CSLAP Cond25
1987	112	115	118	2	ALSC Cond25

DATA SOURCE KEY

CSLAP	New York Citizens Statewide Lake Assessment Program
LCI	the NYSDEC Lake Classification and Inventory Survey conducted during the 1980s and again beginning in 1996 on select sets of lakes, typically 1 to 4x per year
DEC	other water quality data collected by the NYSDEC Divisions of Water and Fish and Wildlife, typically 1 to 2x in any give year
ALSC	the NYSDEC (and other partners) Adirondack Lake Survey Corporation study of more than 1500 Adirondack and Catskill lakes during the mid 1980s, typically 1 to 2x
ELS	USEPA's Eastern Lakes Survey, conducted in the fall of 1982, 1x
NES	USEPA's National Eutrophication Survey, conducted in 1972, 2 to 10x
EMAP	USEPA and US Dept. of Interior's Environmental Monitoring and Assessment Program conducted from 1990 to present, 1 to 2x in four year cycles
Additional data source codes are provided in the individual lake reports	

CSLAP DATA KEY:

The following key defines column headings and parameter results for each sampling season:

Min	Minimum reading for the parameter
Avg	Geometric average (mean) reading for the parameter
Max	Maximum reading for the parameter
N	Number of samples collected
Zsd	Secchi disk transparency, meters
Tot.P	Total Phosphorus as P, in mg/l (Hypo = bottom sample)
NO3	Nitrate + Nitrite nitrogen as N, in mg/l
NH4	Ammonia as N, in mg/l
TDN	Total Dissolved Nitrogen as N, in mg/l
TN	Total Nitrogen as N, in mg/l
TP/TN	Phosphorus/Nitrogen ratios, unitless (calculated from TDN prior to 2006)
Ca	Calcium, in mg/l
Tcolor	True color, as platinum color units
pH	(negative logarithm of hydrogen ion concentration), standard pH
Cond25	Specific conductance corrected to 25°C, in µmho/cm
Chl.a	Chlorophyll a, in µg/l
QA	Survey question re: physical condition of lake: (1) crystal clear; (2) not quite crystal clear; (3) definite algae greenness; (4) high algae levels; and (5) severely high algae levels
QB	Survey question re: aquatic plant populations of lake: (1) none visible; (2) visible underwater; (3) visible at lake surface; (4) dense growth at lake surface; (5) dense growth completely covering the nearshore lake surface
QC	Survey question re: recreational suitability of lake: (1) couldn't be nicer; (2) very minor aesthetic problems but excellent for overall use; (3) slightly impaired; (4) substantially impaired, although lake can be used; (5) recreation impossible
QD	Survey question re: factors affecting answer QC: (1) poor water clarity; (2) excessive weeds; (3) too much algae/odor; (4) lake looks bad; (5) poor weather; (6) litter, surface debris, beached/floating material; (7) too many lake users (boats, jetskis, etc); (8) other

TABLE 1: CSLAP Data Summary for Lake Oscaleta (cont)

Year	Min	Avg	Max	N	Parameter
2006-06	11.7	13.7	15.6	2	CSLAP Ca
2006	11.7	13.7	15.6	2	CSLAP Ca
1987	10.0	10.3	10.6	2	ALSC Ca
Year	Min	Avg	Max	N	Parameter
2006-06	0.16	14.65	53.64	8	CSLAP Chl.a
2006	0.16	14.65	53.64	8	CSLAP Chl.a
2003	5.00	5.50	7.00	5	CedEd Chl.a
Year	Min	Avg	Max	N	Parameter
2006-06	2	2.6	4	8	QA
2006	2	2.6	4	8	QA
Year	Min	Avg	Max	N	Parameter
2006-06	2	2.3	3	8	QB
2006	2	2.3	3	8	QB
Year	Min	Avg	Max	N	Parameter
2006-06	2	2.5	4	8	QC
2006	2	2.5	4	8	QC

- Statistical analyses.** True assessments of water quality trends and comparison to other lakes involve rigid statistical analyses. Such analyses are generally beyond the scope of this program, in part due to limitations on the time available to summarize data from nearly 100 lakes in the five months from data receipt to next sampling season. This may be due in part to the inevitable inter-lake inconsistencies in sampling dates from year to year, and in part to the limited scope of monitoring. Where appropriate, some statistical summaries, utilizing both parametric and non-parametric statistics, have been provided within the report (primarily in Table 2).
- Mean versus Median-** Much of the water quality summary data presented in this report is reported as the **mean**, or the average of all of the readings in the period in question (summer, annual, year to year). However, while mean remains one of the most useful, and often most powerful, ways to estimate the most typical reading for many of the measured water quality indicators, it is a less useful and perhaps misleading estimate when the data are not “normally” distributed (most common readings in the middle of the range of all readings, with readings less common toward the end of the range).

In particular, comparisons of one lake to another, such as comparisons within a particular basin, can be greatly affected by the spread of the data across the range of all readings. For example, the average phosphorus level of nine lakes with very low readings (say 10 µg/l) and one lake with very high readings (say 110 µg/l) could be much higher (in this case, 20 µg/l) than in the “typical lake” in this set of lakes (much closer to 10 µg/l). In this case, **median**, or the middle reading in the range, is probably the most accurate representation of “typical”.

This report will include the use of both mean and median to evaluate “central tendency”, or the most typical reading, for the indicator in question. In most cases, “mean” is used most often to estimate central tendency. However, where noted, “median” may also be used.

**TABLE 2- Present Year and Historical Data Summaries for Lake Osaleta
*Eutrophication Indicators***

Parameter	Year	Minimum	Average	Maximum
Zsd	2006	0.50	2.13	3.10
(meters)	All Years	0.50	2.13	3.10
Parameter	Year	Minimum	Average	Maximum
Phosphorus	2006	0.016	0.031	0.055
(mg/l)	All Years	0.016	0.031	0.055
Parameter	Year	Minimum	Average	Maximum
Chl.a	2006	0.16	14.65	53.64
(µg/l)	All Years	0.16	14.65	53.64

Parameter	Year	Was 2006 Clarity the Highest or Lowest on Record?	Was 2006 a Typical Year?	Trophic Category	Zsd Changing?	% Samples Violating DOH Beach Std?+
Zsd	2006	Not Yet Known	Not Yet Known	Mesotrophic	Not Yet Known	13
(meters)	All Years			Mesotrophic		13
Parameter	Year	Was 2006 TP the Highest or Lowest on Record?	Was 2006 a Typical Year?	Trophic Category	TP Changing?	% Samples Exceeding TP Guidance Value
Phosphorus	2006	Not Yet Known	Not Yet Known	Eutrophic	Not Yet Known	88
(mg/l)	All Years			Eutrophic		88
Parameter	Year	Was 2006 Algae the Highest or Lowest on Record?	Was 2006 a Typical Year?	Trophic Category	Chl.a Changing?	
Chl.a	2006	Not Yet Known	Not Yet Known	Eutrophic	Not Yet Known	
(µg/l)	All Years			Eutrophic		

- + - Minimum allowable water clarity for siting a new NYS swimming beach = 1.2 meters
- + - NYS Total Phosphorus Guidance Value for Class B and Higher Lakes = 0.020 mg/l

The 2006 CSLAP dataset indicates that Lake Osaleta is probably a *eutrophic*, or highly productive lake. Phosphorus readings are consistently high, resulting in high algal densities and reduced water clarity. Although the water in Lake Osaleta is moderately colored, water transparency readings are mostly strongly affected by high algae levels. Water transparency readings are slightly higher than expected given the nutrient and algae levels in the lake. There is only a weak correlation between changes in phosphorus and algae, and between changes in algae and water transparency. However, it is likely that phosphorus readings ultimately control algae levels, which in turn exert the strongest influence on water transparency readings. Deepwater nutrient levels are close to 2x those measured at the lake surface, suggesting that internal nutrient cycling (migration of phosphorus from bottom sediments to bottom waters to surface waters, usually under anoxic conditions) may be significant. Despite this migration, lake productivity does not vary in any predictable way over the course of the summer. Phosphorus levels in Lake Osaleta were above the state guidance value for lakes used for contact recreation (swimming) in all but one of the CSLAP sampling sessions, although Secchi disk transparency readings generally did not fall below the recommended water clarity for swimming beaches (= 1.2 meters). Additional data will help to determine if the 2006 sampling results were typical of Lake Osaleta.

**TABLE 2- Present Year and Historical Data Summaries for Lake Oscaleta
(cont)
Other Water Quality Indicators**

Parameter	Year	Minimum	Average	Maximum
Nitrate	2006	0.01	0.02	0.03
(mg/l)	All Years	0.01	0.02	0.03
Parameter	Year	Minimum	Average	Maximum
NH4	2006	0.01	0.04	0.12
(mg/l)	All Years	0.01	0.04	0.12
Parameter	Year	Minimum	Average	Maximum
TDN	2006			
(mg/l)	All Years			
Parameter	Year	Minimum	Average	Maximum
True Color	2006	12	20	35
(ptu)	All Years	12	20	35
Parameter	Year	Minimum	Average	Maximum
pH	2006	7.39	8.32	9.36
(std units)	All Years	7.39	8.32	9.36
Parameter	Year	Minimum	Average	Maximum
Conductivity	2006	121	133	151
(µmho/cm)	All Years	121	133	151
Parameter	Year	Minimum	Average	Maximum
Calcium	2006	11.7	13.7	15.6
(mg/l)	All Years	11.7	13.7	15.6

These data indicate Lake Oscaleta is a moderately colored, alkaline (above neutral pH) lake with low nitrate and ammonia levels and moderately soft water. Water color readings are probably not high enough to influence water transparency readings, particularly given the oft elevated algae levels in the lake. The color readings suggest moderate levels of dissolved organic matter in the lake, and it is likely that these readings represent “normal” conditions in the lake. The nitrogen to phosphorus ratios suggest that phosphorus controls algae growth. Nitrate and ammonia do not appear to represent a threat to water quality, and readings for both indicators were relatively stable. pH readings are occasionally above the state water quality standards (=6.5 to 8.5), although these readings are probably adequate to support most aquatic organisms. Conductivity readings were somewhat variable over the course of the sampling season, but all readings were indicative of moderately softwater lakes. Calcium levels are above the threshold found to support zebra mussels, but it is not believed that these exotic animals have been found in Lake Oscaleta. Additional data will help to determine if these readings are typical of normal conditions in the lake.

TABLE 2 Present Year and Historical Data Summaries for Lake Oscaleta (cont)

Other Water Quality Indicators (cont)

Parameter	Year	Was 2006 Nitrate the Highest or Lowest on Record?	Was 2006 a Typical Year?	Nitrate High?	Nitrate Changing?	% Samples Exceeding NO3 Standard	
Nitrate	2006	Not Yet Known	Not Yet Known	No	Not Yet Known	0	
(mg/l)	All Years			No		0	
Parameter	Year	Was 2006 Ammonia the Highest or Lowest on Record?	Was 2006 a Typical Year?	NH4 High?	Ammonia Changing?	% Samples Exceeding NH4 Standard	
NH4	2006	Not Yet Known	Not Yet Known	No	Not yet known	0	
(mg/l)	All Years			No		0	
Parameter	Year	Was 2006 TDN the Highest or Lowest on Record?	Was 2006 a Typical Year?	TDN High?	TDN Changing?	Ratios of TN/TP Indicate P or N Limitation?	
TDN	2006	Not Yet Known	Not Yet Known	No	Not yet known	P Limitation	
(mg/l)	All Years			No		P Limitation	
Parameter	Year	Was 2006 Color the Highest or Lowest on Record?	Was 2006 a Typical Year?	Colored Lake?	Color Changing?		
True Color	2006	Not Yet Known	Not Yet Known	No	Not Yet Known		
(ptu)	All Years			No			
Parameter	Year	Was 2006 pH the Highest or Lowest on Record?	Was 2006 a Typical Year?	Acceptable Range?	pH Changing?	% Samples > Upper pH Standard	% Samples < Lower pH Standard
pH	2006	Not Yet Known	Not Yet Known	Yes	Not Yet Known	50	0
(std units)	All Years			Yes		50	0
Parameter	Year	Was 2006 Conductivity Highest or Lowest on Record?	Was 2006 a Typical Year?	Relative Hardness	Conductivity Changing?		
Conductivity	2006	Not Yet Known	Not Yet Known	Intermediate	Not Yet Known		
(µmho/cm)	All Years			Intermediate			
Parameter	Year	Was 2006 Calcium Highest or Lowest on Record?	Was 2006 a Typical Year?	Support Zebra Mussels?	Calcium Changing?		
Calcium	2006	Not Yet Known	Not Yet Known	Yes	Not yet known		
(mg/l)	All Years			Yes			

- + - NYS Nitrate standard = 10 mg/l
- + - NYS Ammonia standard = 2 mg/l (as NH₃-NH₄)
- + - NYS pH standard- 6.5 < acceptable pH < 8.5

TABLE 2- Present Year and Historical Data Summaries for Lake Oscaleta

Lake Perception Indicators (1= most favorable, 5= least favorable)

Parameter	Year	Minimum	Average	Maximum
QA	2006	2	3.1	4
(Clarity)	All Years	2	3.1	4
Parameter	Year	Minimum	Average	Maximum
QB	2006	1	1.4	3
(Plants)	All Years	1	1.4	3
Parameter	Year	Minimum	Average	Maximum
QC	2006	2	2.9	4
(Recreation)	All Years	2	2.9	4

Parameter	Year	Was 2006 Clarity the Highest or Lowest on Record?	Was 2006 a Typical Year?	Clarity Changed?	%Frequency 'Definite Algae Greenness'	%Frequency 'Severe Algae Levels'	%Frequency 'Slightly Impaired' Due to Algae	%Frequency 'Substantially Impaired' Due to Algae
QA	2006	Not yet known	Not yet known	Not yet known	38	25	25	13
(Clarity)	All Years				38	25	25	13
Parameter	Year	Was 2006 Weed Growth the Heaviest on Record?	Was 2006 a Typical Year?	Weeds Changed?	%Frequency Surface Weeds	%Frequency Dense Weeds	%Frequency 'Slightly Impaired' Due to Weeds	%Frequency 'Substantially Impaired' Due to Weeds
QB	2006	Not yet known	Not yet known	Not yet known	25	0	25	0
(Plants)	All Years				25	0	25	0
Parameter	Year	Was 2006 Recreation the Best or Worst on Record?	Was 2006 a Typical Year?	Recreation Changed?	%Frequency Slightly Impaired	%Frequency Substantially Impaired		
QC	2006	Not yet known	Not yet known	Not yet known	38	13		
(Recreation)	All Years				38	13		

The water quality, aquatic plant, and recreational assessments of Lake Oscaleta varied somewhat during the 2006 sampling season. The lake was most frequently described as having “definite algal greenness”, an assessment comparable to that in other lakes with similar water clarity readings. Aquatic plants only occasionally grew to the lake surface, although “excessive weed growth” was occasionally cited as impacting recreational uses of the lake. It is not known whether the offending weeds are exotic or native to the lake. Recreational conditions were most often reported as “slightly” impaired, with use impacts more likely to be associated with poor water clarity and excessive algae growth. Additional data will help to determine if these are representative water quality, plant, and recreational assessments for Lake Oscaleta.

Lake Oscaleta has been described by the CSLAP sampling volunteers as “slightly impaired” during 38% of the CSLAP sampling sessions, and “substantially” impaired 13% of the time. These recreational use impacts were associated with poor water clarity, excessive algae growth, and excessive weed growth, although the most significant impacts corresponded to the lowest water clarity reading in 2006.

How Do the 2006 Data Compare to Historical Data from Lake Oscaleta?

Seasonal Comparison of Eutrophication, Other Water Quality, and Lake Perception Indicators—2006 Sampling Season and in the Typical or Previous Sampling Seasons at Lake Oscaleta

Figures 20 and 21 compare data for the measured eutrophication parameters for Lake Oscaleta in 2006 and since CSLAP sampling began at Lake Oscaleta. Figures 22 and 23 compare nitrogen to phosphorus ratios, Figures 24 through 31 compare other sampling indicators, and Figures 32 and 33 compare volunteer perception responses over the same time periods.

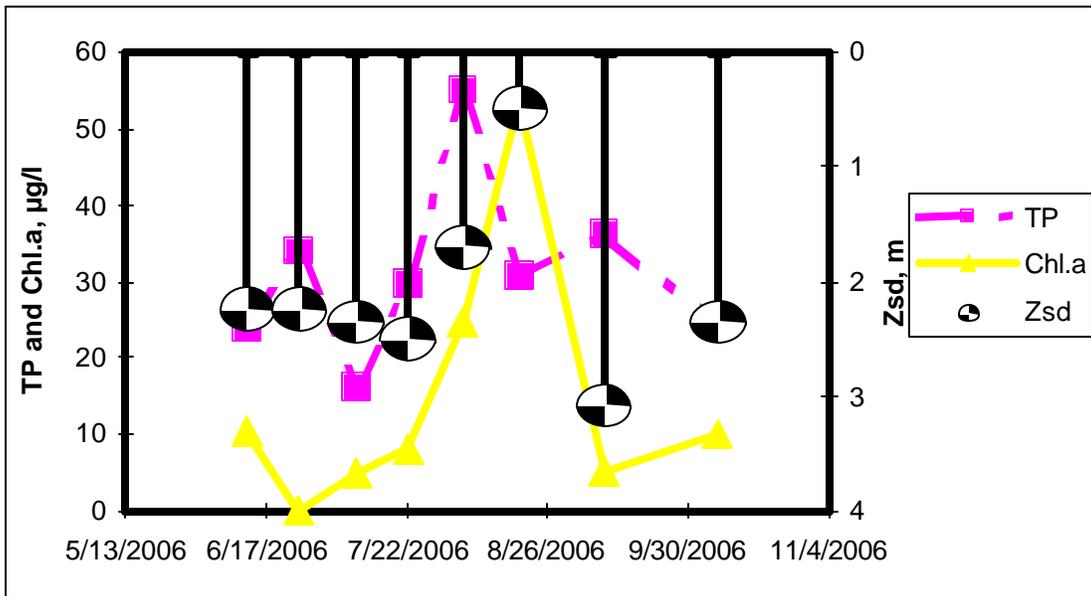


Figure 20. 2006 Eutrophication Data for Lake Oscaleta

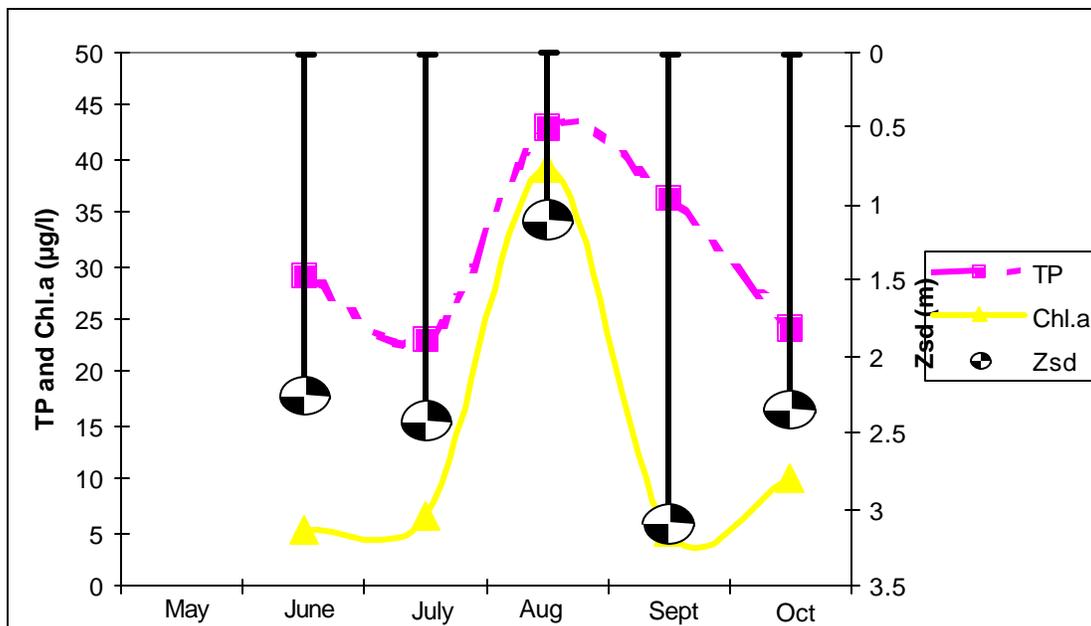


Figure 21- Eutrophication Data in a Typical (Monthly Mean) Year for Lake Oscaleta

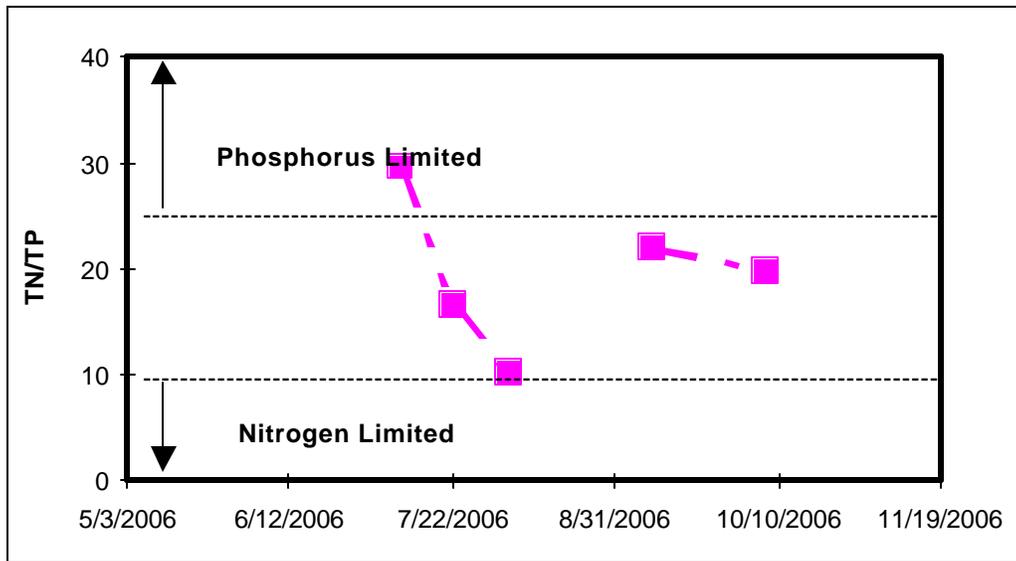


Figure 22. 2006 Nitrogen to Phosphorus Ratios for Lake Osaleta

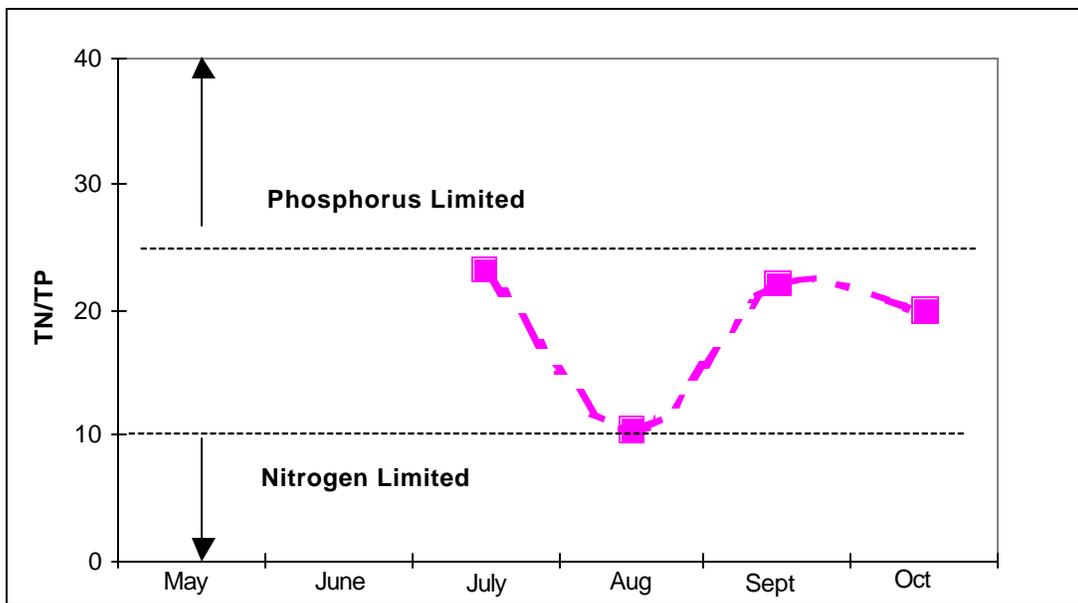


Figure 23- Nitrogen to Phosphorus Ratios in a Typical (Monthly Mean) Year for Lake Osaleta

First Year of CSLAP Sampling
(No Historical Data to Compare)

Figure 24. Annual Average Summer Water Clarity for Lake Oscaleta

First Year of CSLAP Sampling
(No Historical Data to Compare)

Figure 25. Annual Average Summer Chlorophyll a for Lake Oscaleta

First Year of CSLAP Sampling
(No Historical Data to Compare)

Figure 26. Annual Average Summer Total Phosphorus for Lake Oscaleta

First Year of CSLAP Sampling
(No Historical Data to Compare)

Figure 27. Annual Average Summer Total Nitrogen for Lake Oscaleta

First Year of CSLAP Sampling
(No Historical Data to Compare)

Figure 28. Annual Average Summer Nitrate for Lake Oscaleta

First Year of CSLAP Sampling
(No Historical Data to Compare)

Figure 29. Annual Average Summer Ammonia for Lake Oscaleta

First Year of CSLAP Sampling
(No Historical Data to Compare)

Figure 30. Annual Average Summer Conductivity for Lake Oscaleta

First Year of CSLAP Sampling
(No Historical Data to Compare)

Figure 31. Annual Average Summer pH and Color for Lake Oscaleta

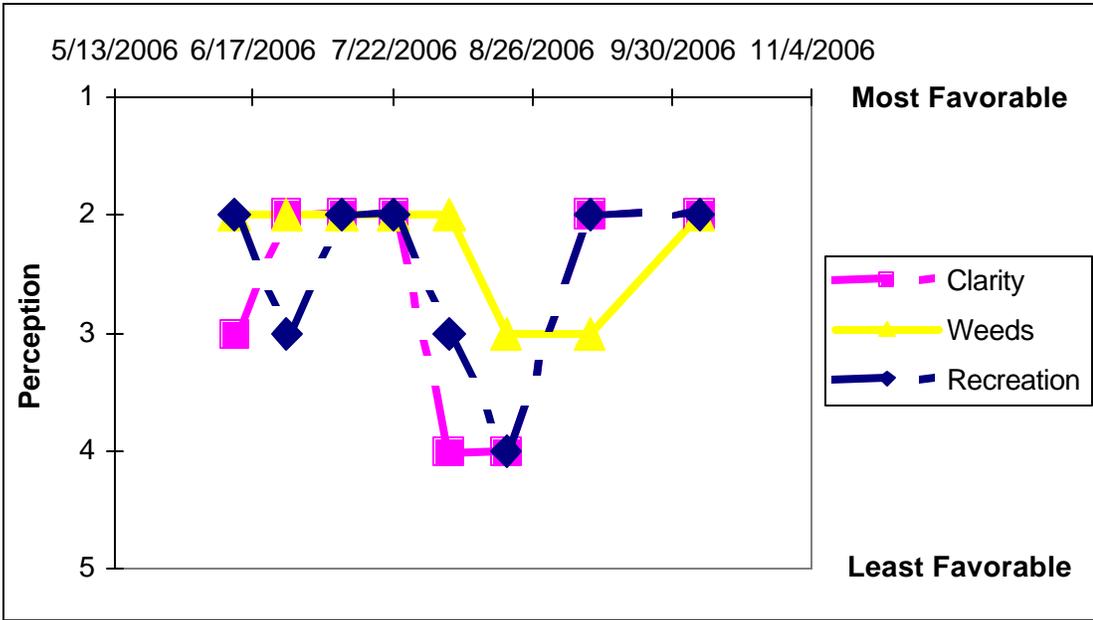


Figure 32. 2006 Lake Perception Data for Lake Oscaleta

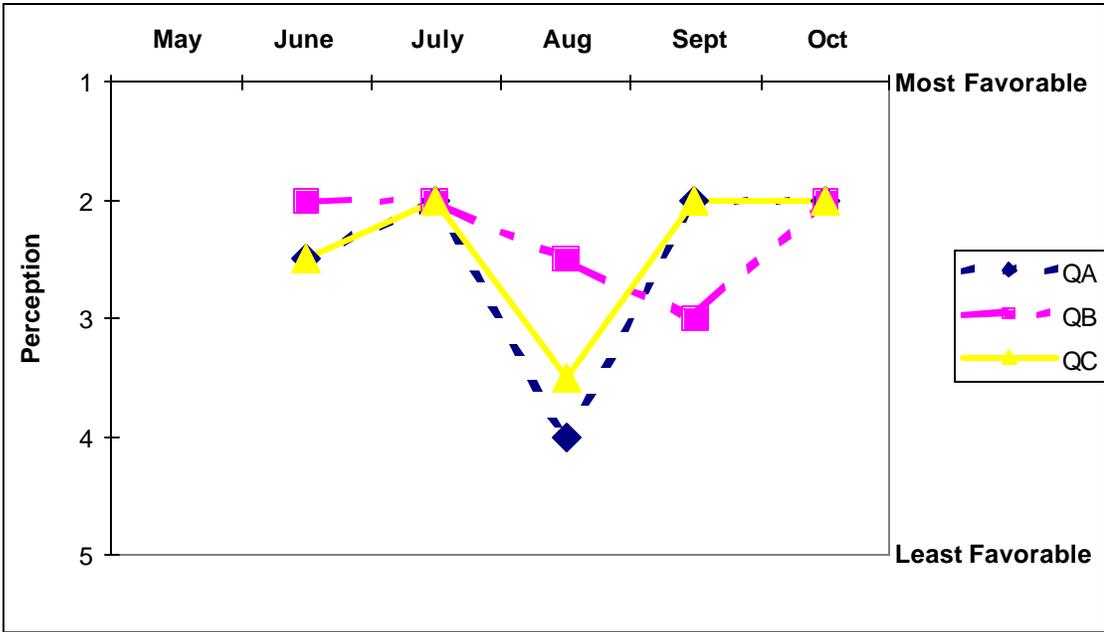


Figure 33- Lake Perception Data in a Typical (Monthly Mean) Year for Lake Oscaleta

First Year of CSLAP Sampling
 (No Historical Data to Compare)

Figure 34- Annual Average Lake Assessments for Lake Oscaleta

(QA = clarity, ranging from (1) crystal clear to (3) definite algae greenness to (5) severely high algae levels
 QB = weeds, ranging from (1) not visible to (3) growing to the surface to (5) dense growth covers lake;
 QC = recreation, ranging from (1) could not be nicer to (3) slightly impaired to (5) lake not usable)

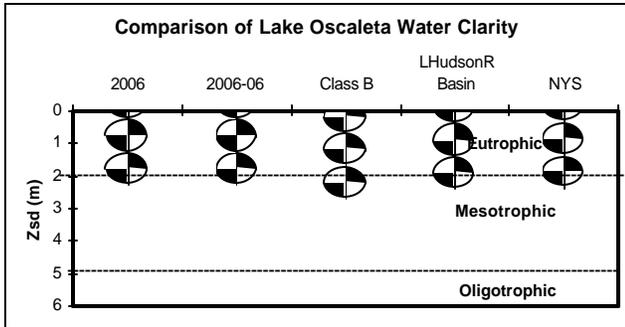


Figure 35. Comparison of 2006 Secchi Disk Transparency to Lakes With the Same Water Quality Classification, Neighboring Lakes, and Other CSLAP Lakes in 2006

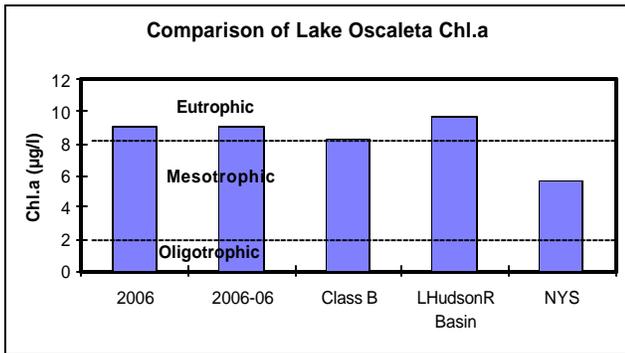


Figure 36. Comparison of 2006 Chlorophyll *a* to Lakes with the Same Water Quality Classification, Neighboring Lakes, and Other CSLAP Lakes in 2006

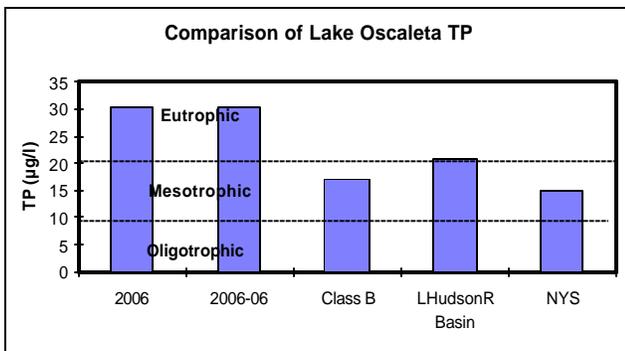


Figure 37. Comparison of 2006 Total Phosphorus to Lakes With the Same Water Quality Classification, Neighboring Lakes, and Other CSLAP Lakes in 2006

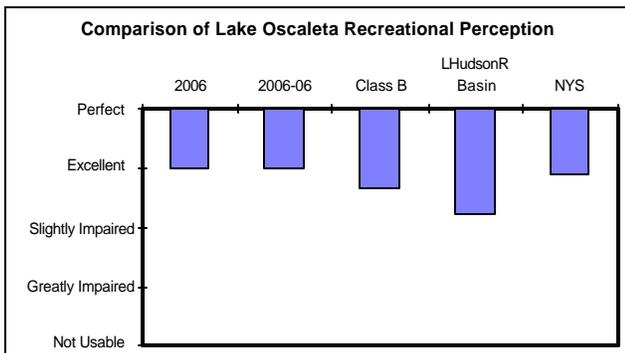


Figure 38. Comparison of 2006 Recreational Perception to Lakes With the Same Water Quality Classification, Neighboring Lakes, and Other CSLAP Lakes in 2006

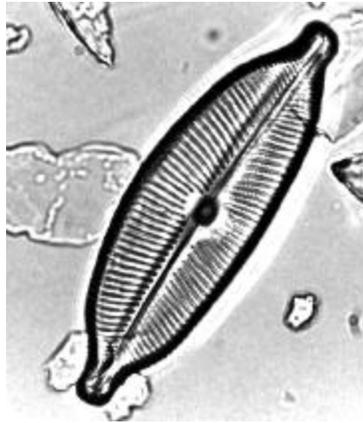
How does Lake Oscaleta compare to other lakes?

Annual Comparison of Median Readings for Eutrophication Parameters and Recreational Assessment For Lake Oscaleta in 2006 to Historical Data for Lake Oscaleta, Neighboring Lakes, Lakes with the Same Lake Classification, and Other CSLAP Lakes

The graphs to the left illustrate comparisons of each eutrophication parameter and recreational perception at Lake Oscaleta-in 2006, other lakes in the same drainage basin, lakes with the same water quality classification (each classification is summarized in Appendix B), and all of CSLAP. Readers should note that differences in watershed types, activities, lake history and other factors may result in differing water quality conditions at your lake relative to other nearby lakes. In addition, the limited data base for some regions of the state preclude a comprehensive comparison to neighboring lakes.

Based on these graphs, the following conclusions can be made about Lake Oscaleta in 2006:

- Using water clarity as an indicator, Lake Oscaleta is slightly more productive than Class B, other Lower Hudson River basin lakes, and other NYS lakes.
- Using chlorophyll *a* concentrations as an indicator, Lake Oscaleta is slightly more productive than other NYS and other Class B lakes, and less productive than other Lower Hudson River basin lakes.
- Using total phosphorus concentrations as an indicator, Lake Oscaleta is more productive than other Lower Hudson River basin, other Class B, and other NYS lakes.
- Using QC on the field observations form as an indicator, Lake Oscaleta is more suitable for recreation than other Lower Hudson River basin, other Class B, and other NYS lakes.



VII. AQUATIC PLANTS

a. Macrophytes:

Aquatic plants should be recognized for their contributions to lake beauty as well as for providing food and shelter for other life in the lake. Emergent and floating plants such as water lilies floating on the lake surface may provide aesthetic appeal with their colorful flowers; sedges and cattails help to prevent shoreline erosion, and may provide food and cover for birds. Submergent plants like pondweeds and leafy waterweed harbor insects, provide nurseries for amphibians and fish, and provide food for birds and other animals. Those who enjoy fishing at the lake appreciate a diverse plant population. Aquatic plants can be found throughout the *littoral zone*, the near-shore areas in which sufficient light reaches the lake bottom to promote photosynthesis. Plant growth in any particular part of the lake is a function of available light, nutrition and space, bottom substrate, wave action, and other factors, and extensive plant growth can occur in both “clean” and “polluted” lakes. A large portion of aquatic vegetation consists of the microscopic algae referred to as phytoplankton; the other portion consists of the larger rooted plants called **macrophytes**.

As invasive plants colonize and spread into a lake, native plant species can be threatened or even eliminated from aquatic plant communities. The most susceptible of these are those that reside in marginal regions, limited by water depth, sediment type, or inability to compete for space. As a result, many plants identified as *rare, threatened or endangered (RTE) species* are protected under New York State law. *Angled spikerush (Eleocharis quadrangulata)* is an *RTE (endangered) species found in Lake Oscaleta in 1999. The plant was also reported by Cedar Eden in 2003.*

Of particular concern to many lakefront residents and recreational users are the *non-indigenous macrophytes* that can frequently dominate native aquatic plants and crowd out more beneficial plant species. The invasive plant species may be introduced to a lake by waterfowl, but in most cases they are introduced by fragments or seedlings that remain on watercraft from already-infested lakes. Once introduced, these species have tenacious survival skills, crowding out, dominating and eventually aggressively overtaking the indigenous (native) plant communities in a variety of water quality conditions. When this occurs, they interfere with recreational activities such as fishing, swimming or water-skiing. These species need to be properly identified to be effectively managed.

Non-native Invasive Macrophyte Species

Examples of the common non-native invasive species found in New York are:

- Eurasian watermilfoil (*Myriophyllum spicatum*)
- Curly-leaf pondweed (*Potamogeton crispus*)
- Eurasian water chestnut (*Trapa natans*)
- Fanwort (*Cabomba caroliniana*).

If these plants are not present, efforts should be made to continue protecting the lake from the introduction of these species.

Whether the role of the lake manager is to better understand the lake ecosystem or better manage the aquatic plant community, knowledge of plant distribution is paramount to the management process. There are many procedures available for assessing and monitoring aquatic vegetation. The CSLAP Sampling Protocol contains procedures for a “semi-quantitative” plant



Figure 39a. *Myriophyllum spicatum* distribution in New York State



Figure 39b. *Potamogeton crispus* distribution in New York State



Figure 39c. *Trapa natans* distribution in New York State



Figure 39d. *Cabomba caroliniana* distribution in New York State

monitoring program. Volunteers collect plant specimens and provide field information and qualitative abundance estimates for an assessment of the macrophyte communities within critical areas of the lake. While these techniques are no substitute for professional plant surveys, they can help provide better information for lake managers. Lake associations planning to devote

significant time and expenditures toward a plant management program are advised to pursue more extensive plant surveying activities.

Formal and informal survey work has been effective in developing statewide distribution maps of each of the major submergent exotic species, and CSLAP data has figured prominently in this process. As of 2006, the statewide distribution maps of confirmed identifications are shown on Figures 39a to 39d:

Aquatic plant surveys have not been conducted through CSLAP at Lake Okauchee. The 1987 ALSC study found the following aquatic plants:

SPECIES	COMMONNAME	SUBM/EMER?	EXOTIC?	DATE	LOCATION
Typha sp	cattail	emergent	no	1987	not reported
Potamogeton sp	pondweed	submergent	?	1987	not reported
Sagittaria sp	arrowhead	emergent	no	1987	not reported
Phragmites sp	reed	emergent	?	1987	not reported
Juncus sp	rush	emergent	no	1987	not reported
Iris sp	iris	emergent	no	1987	not reported
Nuphar sp	yellow waterlily	floating	no	1987	not reported
Nymphaea sp	white waterlily	floating	no	1987	not reported
Brasenia sp	watershield	floating	no	1987	not reported
Hypericum sp	St. Johns wort	submergent?	no	1987	not reported
Myriophyllum sp	milfoil	submergent	?	1987	not reported
Scirpus sp	bulrush	emergent	no	1987	not reported

These plants were not identified down to species level. The 2004 Diagnostic-Feasibility Study of the Lake (conducted by Cedar Eden Environmental, LLC of Saranac Lake, NY) reported that the plant community was dominated by bassweed (Potamogeton amplifolius) and Eurasian watermilfoil (Myriophyllum spicatum). The latter is considered an exotic plant, and may correspond to the milfoil reported in 1987.

b. Algae

Microscopic algae referred to as **phytoplankton** make up much of aquatic vegetation found in lakes. For this reason, and since phytoplankton are the primary producers of food (through photosynthesis) in lakes, they are the most important component of the complex food web that governs ecological interactions in lakes.

In a lake, phytoplankton communities are usually very diverse, and are comprised of hundreds of species having different requirements for nutrients, temperature and light. In many lakes, including those of New York, diatom populations are greatest in the spring, due to a competitive advantage in cooler water and relatively high levels of silica. In most lakes, however, diatom densities rarely reach nuisance portions in the spring. By the summer, green algae take advantage of warmer temperatures and greater amounts of nutrients (particularly nitrogen) in the warm water and often increase in density. These algae often grow in higher densities than do diatoms or most other species, although they are often not the types of algae most frequently implicated in noxious algae blooms. Later in the summer and in the early fall, blue green algae, which possess the ability to utilize atmospheric nitrogen to provide this

required nutrient, increase in response to higher phosphorus concentrations. This often happens right before turnover, or destratification in the fall. These algae are most often associated with taste and odor problems, bloom conditions, and the “spilled paint” slick that prompts the most complaints about algae. Each lake possesses a unique blend of algal communities, often varying in population size from year to year, and with differing species proportional in the entire population. The most common types range from the mentioned diatoms, green, and blue-green algae, to golden-brown algae to dinoflagellates and many others, dominating each lake community.

So how can this be evaluated through CSLAP? CSLAP does assess algal biomass through the chlorophyll *a* measurement. While algal differentiation is important, many CSLAP lake associations are primarily interested in “how much?”, not “what kind?”, and this is assessed through the chlorophyll *a* measurement. Phytoplankton communities have not been regularly identified and monitored through CSLAP, in part due to the cost and difficulty in analyzing samples, and in part due to the difficulty in using a one-time sample to assess long-term variability in lake conditions. A phytoplankton analysis may reflect a temporary, highly unstable and dynamic water quality condition.

Prior to 1998, nearly all CSLAP lakes were sampled once for phytoplankton identification, but since then phytoplankton sampling has not been a regular part of CSLAP. For these sampled lakes, a summary of the most abundant phytoplankton species is included below. Algal species frequently associated with taste and odor problems are specifically noted in this table, although it should be mentioned that these samples, like all other water samples collected through CSLAP, come from near the center of the lake, a location not usually near water intakes or swimming beaches. Since algal communities can also be spatially quite variable, even a preponderance of taste and odor-causing species in the water samples might not necessarily translate to potable water intake or aesthetic impairments, although the threat of such an impairment might be duly noted in the “Considerations” section below.

Phytoplankton surveys have not been conducted through CSLAP at Lake Oscaleta. The Cedar Eden phytoplankton survey determined that the lake was dominated by blue-green algae in early to mid summer, by green algae in late summer, and golden brown algae in the fall, although the early season blue-green algal densities were highest. The corresponding zooplankton surveys found rotifers dominating the early and late summer surveys, with cladoceran dominated the lake in mid summer.

VIII: PRIORITY WATERBODY LISTS AND IMPACTS TO LAKE USE

The Priority Waterbody List (PWL) is presently an inventory of all waters in New York State (lakes, ponds, reservoirs, rivers, streams, and estuaries) known to have designated water uses with some degree of impairment of which are threatened by potential impairment. However, the PWL is slowly evolving into an inventory of all waterbodies for which sufficient information is available to assess the condition and/or usability of the waterbody. PWL waterbodies are identified through a broad network of county and state agencies, with significant public outreach and input, and the list is maintained and compiled by the NYSDEC Division of Water.

Monitoring data from a variety of sources, including CSLAP, have been utilized by state and agencies to evaluate lakes for inclusion on the PWL, and the process for incorporating lakes data has become more standardized.

Specific numeric criteria have recently been developed to characterize sampled lakes in the available use-based PWL categories (*precluded*, *impaired*, *stressed*, or *threatened*). Evaluations utilize the NYS phosphorus guidance value, water quality standards, criteria utilized by other states, and the trophic ranges described earlier to supplement the other more antidotal inputs to the listing. The procedures by which waterbodies are evaluated are known as the Consolidated Assessment and Listing Methodology (CALM) process. This process is undertaken on an annual rotating basin, with waterbodies in several drainage basins evaluated each year. Each of the 17 drainage basins in the state is assessed within every five years. In general, waterbodies that violate pertinent water quality standards (such as those listed in Table 3) at a frequency of greater than 25% are identified as *impaired*, at a frequency of 10-25% are identified as *stressed*, and at a frequency of 0-10% are identified as *threatened*, although some evidence of use impairment (including through CSLAP lake perception surveys) might also be required. Mean (average) phosphorus levels are evaluated against the state guidance value. Evidence of use prohibitions (via beach closures, etc.) is often required to identify a waterbody as *precluded*, while evidence of actual use restrictions or necessary management must accompany an *impaired* listing, at least for lakes evaluated in recent years.

Lakes that have been identified as *precluded* or *impaired* on the PWL are likely candidates for the federal 303(d) list, an “Impaired Waters” designation mandated by the federal Clean Water Act. Lakes on this list must be closely evaluated for the causes and sources of these problems. Remedial measures must be undertaken, under a defined schedule, to solve these water quality problems. This entire evaluation and remediation process is known as the “TMDL” process, which refers to the Total Maximum Daily Load calculations necessary to determine how much (pollution that causes the water quality problems) is too much.

Lake Oscaleta is not presently among the lakes listed on the Lower Hudson River Basin PWL (2001).

TABLE 3- Water Quality Standards Associated With Class B and Higher Lakes

<u>Parameter</u>	<u>Acceptable Level</u>	<u>To Protect....</u>
Secchi Disk Transparency	> 1.2 meters*	Swimming
Total Phosphorus	< 0.020 mg/L and Narrative*	Swimming
Chlorophyll a	none	NA
Nitrate Nitrogen	< 10 mg/L and Narrative*	Drinking Water
Ammonia Nitrogen	2 mg/L*	Drinking Water
True Color	Narrative*	Swimming
pH	< 8.5 and > 6.5*	Aquatic Life
Conductivity	None	NA

*- Narrative Standards and Notes:

Secchi Disk Transparency: The 1.2 meter (4 feet) guidance is applied for safety reasons (to see submerged swimmers or bottom debris), and strictly applies only to citing new swimming beaches, but may be appropriate for all waterbodies used for contact recreation (swimming)

Phosphorus and Nitrogen: “None in amounts that will result in the growths of algae, weeds and slimes that will impair the waters for their best usages” (Class B= swimming)

-The 0.020 mg/l threshold for TP corresponds to a guidance value, not a standard; it strictly applies to Class B and higher waters, but may be appropriate for other waterbodies used for contact recreation (swimming). NYS (and other states) are in the process of identifying numerical nutrient (phosphorus, and perhaps Secchi disk transparency, chlorophyll *a*, and nitrogen) standards, but this is unlikely to be finalized within the next several years.

-The 10 mg/L Nitrate standard strictly applies to only Class A or higher waters, but is included here since some Class B lakes are informally used for potable water intake.

-For the form of ammonia (NH₃+NH₄) analyzed, a 2 mg/l human health standard applies to Class A or higher waters; while lower un-ionized ammonia standards apply to all classes of NYS lakes, this form is not analyzed through CSLAP

Color: “None in amounts that will adversely affect the color or impair the waters for their best usages” (for Class B waters, this is swimming)

pH: The standard applies to all classes of waterbodies

1. Water Quality Standards Evaluation on Lake Oscaleta:

pH readings exceeded the NYS water quality standards (=6.5 to 8.5) during half of the CSLAP sampling sessions at Lake Oscaleta. Phosphorus levels at Lake Oscaleta exceeded the phosphorus guidance value for NYS lakes (=0.020 mg/l) during all but one of the CSLAP sampling sessions, but water transparency readings failed to reach the minimum recommended water clarity for swimming beaches (= 1.2 meters) during only one of the CSLAP sampling sessions. It is not known if any of the narrative water quality standards listed in Table 3 have been violated at Lake Oscaleta; none of the other numeric standards listed in Table 3 have been violated.

2. Lake Uses:

Water quality monitoring programs are devised to evaluate lake conditions as they relate to a variety of lake indicators, from water quality standards to trophic conditions to invasive species to other measures of the physical, chemical, and biological integrity of these ecological systems. One of these indicators is intended to be lake uses- whether these lakes and ponds can be used for potable water, swimming and bathing, fishing and use of the water by aquatic life, and aesthetics. This is consistent with the broad goals of the 1972 federal Clean Water Act, the governing legislation for federal and state management of lakes and ponds, which state that a fundamental goal of environmental management was to make all waterbodies “fishable and swimmable” by 1983.

The “fishability” of a lake or pond is a function of water quality (are there pollutants that will kill the fish or render them inedible?), substrate and habitat (is there enough cold water and high oxygen for coldwater fish?, is there enough food for the fish? is there enough cover from predators or structure for fishermen?), space (is there enough flowing water for survival or reproduction?, is there enough room to support all of the various fish species in the lake?), and even access (can anglers get to the areas where the fish can be found?).

Likewise, the “swimmability” of a lake or pond also depends on water quality (will I get sick due to bacterial contamination from sewage, stormwater or waterfowl?), safety (can swimmers or bottom debris be seen in deeper water?), aesthetics (is the water too green, too weedy, or too cold?, is the bottom too mucky?), user conflicts (can I swim where people jet ski?), the physical characteristics of the lake and shoreline (how quickly does the lake get too deep? Is the shoreline flat enough for a beach?), legal considerations (will the threat of litigation prevent a lake community from establishing public beaches?), and also access (can swimmers from less hospitable parts of the lake or from the outside swim at a beach?).

Although other designated lake uses are not identified as primary goals of the Clean Water Act, they should be evaluated as part of the lake assessment process. These include potable water, non-contact recreational uses such as boating, aquatic life support unrelated to fishing, and aesthetics. Similar questions could be posed about the suitability of a particular lake or pond for this use, although many of the concerns addressed in evaluating the fishability or swimmability of a waterbody are pertinent to evaluating drinking water quality, the ability of a lake to support power boating or sailing, or the adequacy of the lake bottom for salamanders, frogs, and other valued biota.

CSLAP is not really designated to answer many of these questions, at least directly. Some of these issues relate to the physical characteristics of the entire shoreline and bottom of the lake or pond, and cannot be easily evaluated in simple water quality surveys. Other important water quality indicators, such as bacteria, cannot be sampled at the frequency needed to compare lake conditions to existing water quality standards, or are limited by logistic considerations. Other indicators, such as sediment toxins, are too expensive to be included in standard water quality monitoring programs. It is anticipated that future generations of CSLAP will look to better address some of these questions, through expanded monitoring and partnerships with other monitoring agencies, academic institutions, lake residents, and other parties invested in the lake assessment and management process. It is also anticipated that data from other sources will be more completely included in the lake and pond assessment process in the future. Until that time, however, it should again be stated that these assessments are both preliminary and incomplete, based on data presently collectable through the monitoring programs summarized in this report.

Lake Oscaleta is a Class B lake, which means it is designated for support of contact recreation (bathing and swimming), aquatic life (including fishing), non-contact recreation (such as boating) and aesthetics. It is not designated for potable water (drinking) use, although some lake residents may use lake water for this purpose. As such, Lake Oscaleta should be evaluated for its best intended uses- contact recreation, support of aquatic life, non-contact recreation, and aesthetics.

a. Potable Water

As noted above, Lake Oscaleta is not classified for potable water use.

b. Swimming/Contact Recreation

It is presumed that Lake Oscaleta is used for swimming, bathing, or other forms of contact recreation, although the frequency of and opportunities for swimming are not evaluated through CSLAP. As noted above, it is classified for bathing and swimming.

There are a number of water quality indicators measured in CSLAP that relate to the suitability of lake or pond for swimming and contact recreation. Water clarity measurements can be used to evaluate the lake against the NYS Department of Health guidelines for siting new swimming beaches (= 4 feet). Public perception data collected through CSLAP assess swimming conditions, and regional or statewide criteria connecting water transparency readings (or nutrient and algae levels) to recreational use impacts will likely be developed in the near future. However, there remains a relatively strong correlation between contact recreational conditions and phosphorus readings, with recreational use impacts generally corresponding to the state guidance value for phosphorus (= 20 parts per billion total phosphorus). Algae levels are measured as chlorophyll *a*, while rooted aquatic plant populations are broadly quantified through CSLAP, and are linked to potential impacts on swimming and aesthetics. These water quality-based and perception-based evaluations of swimming conditions are outlined below.

1. Water Quality Evaluation of Swimming/Contact Recreation

These data showed that all but one (88%) of the Lake Oscaleta samples possessed total phosphorus readings exceeding 20 parts per billion (=µg/l), which corresponds to the state phosphorus guidance value. Water transparency readings were less than 2 meters during 25% of the CSLAP sampling sessions. This roughly corresponds to the distinction between *eutrophic* and *mesotrophic* lakes and a water clarity reading that would roughly be equivalent to the state phosphorus guidance value. Perhaps more importantly, this may correspond to the saddle point between high quality and reduced quality swimming, based on lake perception data (see below).

Although there is no state water quality standard for chlorophyll *a*, readings exceeding 8 µg/l generally correspond to water clarity readings lower than 2 meters and total phosphorus readings in excess of 20 µg/l- each of these indicator thresholds marks the distinction between *mesotrophic* and *eutrophic* lake. 63% of the Lake Oscaleta samples corresponded to chlorophyll *a* readings > 8 µg/l.

Bacteria data have not been collected through CSLAP on Lake Oscaleta, and (if collected by the lake association or local community) have not been forwarded to the NYSDEC for evaluation.

2. Lake Perception Evaluation

Lake perception data from CSLAP provide insights into recreational (swimming) conditions, perceptions of water clarity, and the density and coverage of aquatic plants. Recreational assessments indicating “beautiful, could not be nicer” and “..excellent for swimming, boating, and overall enjoyment” conditions suggest no limits to recreational use. The

frequency of “slightly” to “substantially” impaired conditions may be closely related to the need to implement lake management actions. These surveys also assess the extent to which these impacts are influenced by excessive weed growth, nuisance algae or poor water clarity.

The evaluation of these survey results, and the extrapolation of these results to a Lake-wide assessment, are restricted by the small sample size, and the potential for responses that are not representative of the responses from the typical lake resident, whether due to the impact of local conditions or different goals for different lake users. However, these assessments may serve as an instructive starting point for evaluating impacts on lake uses.

The CSLAP volunteers reported that Lake Oscaleta was described as “slightly impaired” during 38% of the sampling sessions, and “substantially impaired” 13% of the time. Slightly impaired conditions were associated with “poor water clarity” or “excessive algae growth” 25% of the time, and with “excessive weed growth” 25% of the time, while “substantially impaired” conditions were associated with poor water clarity and excessive algae 13% of the time.

3. Overall Evaluation- Swimming and Contact Recreation

The CSLAP dataset at Lake Oscaleta, including water chemistry data, physical measurements, and volunteer samplers’ perception data, was too limited to evaluate use impairments. Additional data may help to provide initial evaluations in coming years.

c. Aquatic Life/Non-Contact Recreation

Lake Oscaleta supports fishing and other forms of non-contact recreation. Other forms of non-contact recreation, such as boating, may be a function of access points, whether the lake shoreline is inhabited, and water depth, but it is also presumed that Lake Oscaleta may be used for boating.

While water quality plays a role in evaluating non-contact recreation, particularly cold water fisheries, the information needed to properly evaluate fishing quality, angler success, and boating enjoyment and viability are not collected in most routine monitoring programs. It is anticipated that future generations of the CSLAP report will include more comprehensive evaluations of non-contact recreational conditions in lakes and ponds in the lake, as databases containing this information become more readily available, but until that time, only ancillary measures can be evaluated.

The primary indicators from these monitoring programs used to evaluate fisheries, aquatic life, and non-contact recreation (boating, etc.) include lake perception surveys, aquatic plant densities (and the presence of invasive exotic plants), and water quality indicators related to fish habitat and survival, such as pH and ammonia. While other water quality indicators, such as other forms of nitrogen, can also be used to evaluate water quality impacts to aquatic life, these indicators are generally found at low enough levels to minimize their utility in evaluating lake conditions. Dissolved oxygen can be very useful in evaluating habitat, and oxygen depth profiles have been evaluated through CSLAP (see Appendix F). These datasets can provide at least some insights into the ability of lakes and ponds to support these uses.

1. Fisheries and Aquatic Life Evaluation

pH data are collected through CSLAP. Fish consumption advisories are issued by the NYS Department of Health, and fishing regulations are instituted by the NYSDEC. Lake recreational perception data related to non-contact recreation (fishing and boating) and aesthetics are also collected through CSLAP, and these can be used to evaluate fisheries and aquatic life impacts to Lake Osaleta.

These data indicate that pH readings in half of the Lake Osaleta samples exceeded the state water quality standards (= 6.5 to 8.5). While laboratory pH is not as accurate as field pH for evaluating lake acidity, these data suggest that no fisheries or aquatic life impacts come from low pH. It is not known if high pH impacts the fisheries- CSLAP data are not adequate to evaluate fisheries health

Dissolved oxygen data have been collected through CSLAP, and through the ALSC and Cedar Eden projects. The CSLAP data show that the lake is thermally stratified below a depth of about 5 meters, and anoxia occurs below a depth of about 6 meters. The ALSC data identified hypoxia (depressed oxygen readings) near the bottom, while the Cedar Eden data found more significant deepwater oxygen depression (readings below 1 ppm) throughout the summer near the lake bottom.

The list of fish species provided in the ALSC report does not indicate any species susceptible to low temperature and low dissolved oxygen conditions. It should also be noted that there are no specific fish consumption advisories on Lake Osaleta.

2. Boating (Recreation) and Aesthetics Evaluation

Impacts to non-contact recreation, such as boating and aesthetics, can only be peripherally evaluated through CSLAP. Sampling volunteers can report that the lake “looks bad”, as a direct measure of impacts to lake aesthetics, while “poor water clarity”, “excessive algae growth”, and “excessive weed growth” may be indirect measures of these impacts.

The CSLAP volunteers reported that Lake Osaleta “looks bad” during 25% of the sampling sessions (in late summer), and surface weed growth was reported during 25% of the CSLAP sessions (late August to early September).

3. Overall Evaluation- Aquatic Life and Non-Contact Recreation

The CSLAP dataset on Lake Osaleta, including water chemistry data, physical measurements, and volunteer samplers’ perception data, are insufficient to evaluate the aquatic life and non-contact recreational impacts on the lake. Additional data may help to provide initial evaluations in coming years.

IX: CONSIDERATIONS FOR LAKE MANAGEMENT

CSLAP is intended for a variety of uses, such as collecting needed information for comprehensive lake management, although it is not capable of collecting all the needed information. To this end, this section includes a *broad summary of the major lake problems and “considerations” for lake management*. These include only those lake problems which may have been defined by CSLAP sampling, such as physical condition (algae and water clarity), aquatic plant coverage (type and extent of weed populations), and recreational suitability of the lake, as related to contact recreation. These broad categories may not encompass the most pressing issue at a particular time at any given CSLAP lake; for example, local concerns about filamentous algae or concerns about other parameters not analyzed in the CSLAP sampling. While there is some opportunity for CSLAP trained volunteers to report and assess some site-specific conditions or concerns on the CSLAP Field Observations Form, such as algae blooms or shoreline vegetation, this section is limited to the confines of this program. The categories represent the most common, broadest issues within the lake management as reported through CSLAP.

Each summarized management strategy is more extensively outlined in Diet for a Small Lake, and this joint NYSDEC-NYSFLA publication should be consulted for more details and for a broader context of in-lake or watershed management techniques. These “considerations” should not be construed as “recommendations”, since there is insufficient information available through CSLAP to assess if or how a lake should be managed. Issues associated with local environmental sensitivity, permits, and broad community management objectives also cannot be addressed here. Rather, the following section should be considered as “tips” or a compilation of suggestions for a lake association to manage problems defined by CSLAP water quality data or articulated by perception data. When appropriate, lake-specific management information, and other lake-specific or local “data” (such as the presence of a controllable outlet structure) is reported in *bold* in this “considerations” section.

The primary focus of CSLAP monitoring is to evaluate lake condition and impacts associated with lake eutrophication. Since lake eutrophication is often manifested in excessive plant growth, whether algae or aquatic macrophytes (weeds), it is likely that lake management activities, whether promulgated to reduce algae or weed growth, or to maintain water clarity and the existing makeup and density of aquatic plants in the lake, will need to address watershed inputs of nutrients and sediment to the lake, since both can contribute to either algal blooms or excessive weed growth. A core group of nutrient and sediment control activities will likely serve as the foundation for most comprehensive lake management plans and activities, and can be summarized below.

a. GENERAL CONSIDERATIONS FOR ALL CSLAP LAKES

Nutrient controls can take several forms, depending on the original source of the nutrients:

- Septic systems can be regularly pumped or upgraded to reduce the stress on the leach fields which can be replaced with new soil or moving the discharge from the septic tank to a new field). Pumpout programs are usually quite inexpensive, particularly when lakefront residents negotiate a bulk rate discount with local pumping companies. Upgrading systems can be

expensive, but may be necessary to handle the increased loading from camp expansion or conversion to year-round residency. Replacing leach fields alone can be expensive and limited by local soil or slope conditions, but may be the only way to reduce actual nutrient loading from septic systems to the lake. It should be noted that upgrading or replacing the leach field may do little to change any bacterial loading to the lake, since bacteria are controlled primarily within the septic tank, not the leach field.

- Stormwater runoff control plans include street cleaning, artificial marshes, sedimentation basins, runoff conveyance systems, and other strategies aimed at minimizing or intercepting pollutant discharge from impervious surfaces. The NYSDEC has developed a guide called Reducing the Impacts of Stormwater Runoff to provide more detailed information about developing a stormwater management plan. This is a strategy that cannot generally be tackled by an individual homeowner, but rather requires the effort and cooperation of lake residents and municipal officials.
- There are numerous agriculture management practices such as fertilizer controls, soil erosion practices, and control of animal wastes, which either reduce nutrient export or retain particles lost from agricultural fields. These practices are frequently employed in cooperation with county Soil and Water Conservation District offices, and are described in greater detail in the NYSDEC's Controlling Agricultural Nonpoint Source Water Pollution in New York State. Like stormwater controls, these require the cooperation of many watershed partners, including farmers.
- Streambank erosion can be caused by increased flow due to poorly managed urban areas, agricultural fields, construction sites, and deforested areas, or it may simply come from repetitive flow over disturbed streambanks. Control strategies may involve streambank stabilization, detention basins, revegetation, and water diversion.

Land use restrictions development and zoning tools such as floodplain management, master planning to allow for development clusters in more tolerant areas in the watershed and protection of more sensitive areas; deed or contracts which limit access to the lake, and cutting restrictions can be used to reduce pollutant loading to lakes. This approach varies greatly from one community to the next and frequently involves balancing lake use protection with land use restrictions. State law gives great latitude to local government in developing land use plans.

Lawn fertilizers frequently contain phosphorus, even though nitrogen is more likely to be the limiting nutrient for grasses and other terrestrial plants. By using lawn fertilizers with little or no phosphorus, eliminating lawn fertilizers or using lake water as a “fertilizer” at shoreline properties, fewer nutrients may enter the lake. Retaining the original flora as much as possible, or planting a buffer strip (trees, bushes, shrubs) along the shoreline, can reduce the nutrient load leaving a residential lawn.

Waterfowl introduce nutrients, plant fragments, and bacteria to the lake water through their feces. Feeding the waterfowl encourages congregation which in turn concentrates and increases this nutrient source, and will increase the likelihood that plant fragments, particularly from Eurasian watermilfoil and other plants that easily fragment and reproduce through small fragments, can be introduced to a previously uncolonized lake.

Although not really a “watershed control strategy”, establishing **no-wake zones** can reduce shoreline erosion and local turbidity. Wave action, which can disturb flocculent bottom sediments and unconsolidated shoreline terrain is ultimately reduced, minimizing the spread of fertile soils to susceptible portions of the lake.

Do not discard or introduce plants from one water source to another, or deliberately introduce a "new" species from catalogue or vendor. For example, do not empty bilge or bait bucket water from another lake upon arrival at another lake, for this may contain traces of exotic plants or animals. Do not empty aquaria wastewater or plants to the lake.

Boat propellers are a major mode of transport to uncolonized lakes. Propellers, hitches, and trailers frequently get entangled by weeds and weed fragments. Boats not cleaned of fragments after leaving a colonized lake may introduce plant fragments to another location. New introductions of plants are often found near public access sites.

b. SPECIFIC CONSIDERATIONS FOR LAKE OSCALETA

Management Focus: **Water Clarity/Algae/Physical Condition/Recreational Condition**

Problem	Probable cause	Probable source
Poor water clarity	Excessive algae	Excessive phosphorus loading from septics, watershed runoff (stormwater, construction sites, agriculture, ...)

Discussion:

The water sampling results indicate that recreational impairments in this lake are related to lower-than-desired water transparency. Water clarity in this lake appears to be strongly related to (planktonic) algae, which is linked to nutrient concentrations. A management focus to improve water clarity involves reducing algae levels, which is linked to reducing nutrient concentrations in the lake and within the watershed. These considerations do not constitute recommendations, since it is not known if the lake association is attempting to improve water clarity, but these considerations are a discussion of some management alternatives which may have varying levels of success addressing these problems.

POTENTIAL IN-LAKE CONTROLS

The strategies outlined below primarily address the cause, but not the ultimate source, of problems related to poor water clarity. As such, their effectiveness is necessarily short-term, but perhaps more immediately realized, relative to strategies that control the source of the problem. The problems may continue or worsen if the source of the problem, excessive nutrients, is not addressed, using strategies such as those described under **Watershed Controls** below. In-lake controls are listed in order of frequency of use in the “typical” NYS lake: *copper sulfate, precipitation/inactivation, hypolimnetic withdrawal, aeration, dilution/flushing, artificial circulation, and food web manipulation.*

- *Copper sulfate* is an algacide that is frequently used to control nuisance levels of planktonic algae (dots of algae throughout the water column) or filamentous algae

(mats of algae on the lake surface, weeds, or rocks) throughout the lake. It is usually applied 1-3x per summer in granular or liquid form, usually by a licensed applicator. Many people feel that it is effective at reducing algae levels to below nuisance conditions, others feel it only “flattens the peak” of the worst blooms, and still others think it is merely a placebo, given the short – lived dominance of some phytoplankton species. There are concerns about the long-term affect of copper on the lake bottom, including the effects on bottom macroinvertebrate communities, and implications of increasing the concentrations of copper as a component of bottom sediments. Another concern is a possible deleterious affect of copper on the zooplankton (microscopic animals that feed on algae) community, which could, in some lakes, ultimately cause a “bounce-back” algae bloom that is worse than the original bloom. *It is not believed by the report authors that copper sulfate has been used at Lake Oscaleta.*

- *Precipitation/Inactivation* involves adding a chemical binding agent, usually alum, to bind and precipitate phosphorus, removing it from the water column, and to seal bound phosphorus in the sediment, rendering it inactive for release to the overlying water (as often occurs in stratified lakes with low oxygen levels). It has a mixed rate of success in NYS, although when successful it usually provides long-term control of nutrient release from bottom sediments (it is only a short-term method for removing existing phosphorus from the water column). It is not recommended for lakes with low pH or buffering capacity (like most small NYS lakes at high elevation), for at low pH, aluminum can be toxic to fish. Since CSLAP does not conduct extensive deepwater monitoring, or any sediment release rate studies, the efficacy of this strategy, based on CSLAP data, is not known. *It is not believed that this has been attempted at Lake Oscaleta.*
- *Hypolimnetic withdrawal* takes deoxygenated, high nutrient water from the lake bottom and discharges the water downstream from the lake. This strategy is sort of a hybrid of aeration and dilution/flushing, and is usually limited to lakes in which control structure (such as a dam) exists where the release valve is located below the thermocline. It has been quite successful and usually inexpensive when applied properly, but must only be employed when downstream waterbodies will not be adversely impacted by the pulse of low oxygen water (which may include elevated levels of hydrogen sulfide, ammonia, and iron). *It is not known by the report authors if this is feasible at Lake Oscaleta.*
- *Aeration* involves pumping or lifting water from the lake bottom (hypolimnion) for exposure to the atmosphere, with the oxygenated waters returning to the lake bottom. The airlift device is usually quite expensive, and operating costs can be quite high. There is also a risk of breaking down the thermocline, which can result in an increase in algae levels and loss of fish habitat for many cold-water species. However, most of the limited number of aeration projects has been quite successful. Since CSLAP does not collect dissolved oxygen data for most program lakes, it is not definitively known whether aeration (or hypolimnetic withdrawal) would benefit this lake. *Artificial circulation* is the process by which air is injected into the hypolimnion to eliminate thermal stratification- it is aeration by circulation.

- *Dilution/flushing* involves using high quality dilution water to reduce the concentration of limiting nutrients and increase the rate at which these nutrients are flushed through the lake. This strategy requires the availability of high quality dilution water and works best when the lake is small, eutrophic, and no downstream waterbodies that may be affected by the pulse of nutrients leaving the lake. For many lakes, however, high quality dilution water is probably not available from the surrounding watershed, because such an input would already be flushing the lake. *This is probably the case with Lake Oscaleta.*
- *Food web manipulation* involves altering the population of one component within the food web, most frequently algae, by altering the populations of other components in the same web. For algae control, this would most frequently involve stocking the lake with herbivorous (algae-eating) fish, but this may be at the expense of other native fish. While this procedure has worked in some situations, as with most attempts at biomanipulation, altering the food chain may be risky to the whole ecosystem, and not recommended at lakes in which the native fisheries serve as a valuable local resource.

Management Focus: **The Impact of Weeds on Recreational Condition**

Issue	Effect on Lake Use
Low weed growth	No use impairments associated with weed growth

Discussion:

Weed growth in this lake is not dense enough to have an impact on recreational or aesthetic quality of the lake- perhaps due to active control (see below). For many lake associations, this is the ideal situation, although an ideal condition for swimmers, boaters and lakefront residents may not be ideal for a significant sports fishery. For lakes in this condition, lake management is largely a task of maintaining course, of keeping nuisance plants out of the lake.

-If you have a small amount of nuisance plant growth you may want to consider the following-

-Hand harvesting is a very labor-intensive means for controlling weed populations. If only a very small number of nuisance plant stems exist, this may be the best means of control, removing the roots and stems of the entire plant, and disposing properly before they propagate into larger, uncontrollable beds that become the obnoxious neighbors of beneficial native plants.

-Benthic barriers are small opaque mats (usually constructed from plastic, burlap, or other materials) anchored down on top of plants to prevent sunlight from reaching the plants, thus eventually killing the plants. These are limited to only small areas, and the mats must be anchored and perforated to prevent gas bubbles from dislodging the mats.

c. SPECIFIC MONITORING CONSIDERATIONS FOR LAKE OSCALETA

Discussion:

Lake Oscaleta was sampled through CSLAP for the first time in 2006. More extensive data will help to evaluate “normal” conditions on the lake, and to identify water quality or use problems at the lake. However, some additional parameters may be appropriate for evaluation at the lake:

1. *Bacteria*- Lake Oscaleta is classified for use for contact recreation (swimming), and it is likely that some swimming occurs. The use of the lake for swimming and bathing can best be evaluated with bacteriological data. The state water quality standards reference sampling schedules requiring at least five samples per month. These data cannot be collected through CSLAP.
2. *Algal toxins*- Algal toxins, usually associated with blue green algae, may affect swimmers and others who ingest small amounts of water (as well as any lake residents who utilize Lake Oscaleta as a potable water supply, despite the lake classification). These may be analyzed in standard water samples as part of CSLAP in coming years.
3. *Aquatic plants*- Aquatic plant surveys have not been conducted through CSLAP at Lake Oscaleta. CSLAP samplers can collect and submit for identification any plant samples thought to be exotic or otherwise invasive, as well as any rare or unusual plants. Sampling protocols are also available to conduct systematic monitoring of aquatic plants for the purpose of evaluating aquatic plant management actions utilized at the lake. This is particularly important given the increasing concern about exotic plant growth in the Lower Hudson River region of the state.
4. *Temperature and oxygen profiles*- the suitability of the lake for supporting sensitive fish, the susceptibility of the lake to nutrient release from bottom sediments and fall algal blooms, and the environment for aquatic plant growth can be evaluated through temperature and oxygen profiles. These can be created through the use of electronic meters or through chemical titrations conducted on site, and these data should continue to be collected through CSLAP.

Appendix A. Raw Data for Lake Oscaleta

LNum	LName	Date	Zbot	Zsd	Zsamp	Tot.P	NO3	NH4	TDN	TN/TP	TColor	pH	Cond25	Ca	Chl.a
205	L Oscaleta	6/12/2006	10.6	2.26	1.5	0.024	0.03	0.04			17	7.78	135	15.6	10.46
205	L Oscaleta	6/25/2006	10.7	2.25	1.5	0.034	0.02	0.02			12	7.62	128		0.16
205	L Oscaleta	7/9/2006	10.5	2.35	1.5	0.016	0.01	0.06	0.48	29.63	15	8.89	121		4.97
205	L Oscaleta	7/22/2006	11.0	2.50	1.5	0.030	0.01	0.01	0.50	16.68	21	8.70	133		8.11
205	L Oscaleta	8/5/2006	10.8	1.70	1.5	0.055	0.02	0.03	0.57	10.32	35	9.11	131	11.7	24.62
205	L Oscaleta	8/19/2006	10.6	0.50	1.5	0.031	0.01	0.01			24	9.36	124		53.64
205	L Oscaleta	9/9/2006	10.7	3.10	1.5	0.036	0.02	0.12	0.80	22.02	18	7.68	138		5.17
205	L Oscaleta	10/7/2006	10.8	2.35	1.5	0.024	0.01	0.05	0.48	19.99	15	7.39	151		10.07
205	L Oscaleta	6/12/2006	10.6												
205	L Oscaleta	6/25/2006	10.7		9.2	0.038									
205	L Oscaleta	7/9/2006	10.5		9.0	0.044									
205	L Oscaleta	7/22/2006	11.0		9.5	0.046									
205	L Oscaleta	8/5/2006	10.8		9.0	0.051									
205	L Oscaleta	8/19/2006	10.6		9.0	0.082									
205	L Oscaleta	9/9/2006	10.7		9.0	0.077									
205	L Oscaleta	10/7/2006	10.8		9.5	0.061									

LNum	LName	Date	TAir	TH2O	QA	QB	QC	QD
205	L Oscaleta	6/12/2006	23	20	3	2	2	0
205	L Oscaleta	6/25/2006	23	24	2	2	3	2
205	L Oscaleta	7/9/2006	27	25	2	2	2	12
205	L Oscaleta	7/22/2006	25	27	2	2	2	125
205	L Oscaleta	8/5/2006	31	31	4	2	3	1234
205	L Oscaleta	8/19/2006	32	27	4	3	4	134
205	L Oscaleta	9/9/2006	29	23	2	3	2	12
205	L Oscaleta	10/7/2006	15	17	2	2	2	25
205	L Oscaleta	6/12/2006						
205	L Oscaleta	6/25/2006						
205	L Oscaleta	7/9/2006						
205	L Oscaleta	7/22/2006		8				
205	L Oscaleta	8/5/2006		8				
205	L Oscaleta	8/19/2006		8				
205	L Oscaleta	9/9/2006		8				
205	L Oscaleta	10/7/2006		9				

Appendix B. New York State Water Quality Classifications

- Class N: Enjoyment of water in its natural condition and where compatible, as source of water for drinking or culinary purposes, bathing, fishing and fish propagation, recreation and any other usages except for the discharge of sewage, industrial wastes or other wastes or any sewage or waste effluent not having filtration resulting from at least 200 feet of lateral travel through unconsolidated earth. These waters should contain no deleterious substances, hydrocarbons or substances that would contribute to eutrophication, nor shall they receive surface runoff containing any such substance.
- Class AA_{special}: Source of water supply for drinking, culinary or food processing purposes; primary and secondary contact recreation; and fishing. These waters shall be suitable for fish propagation and survival, and shall contain no floating solids, settleable solids, oils, sludge deposits, toxic wastes, deleterious substances, colored or other wastes or heated liquids attributable to sewage, industrial wastes or other wastes. There shall be no discharge or disposal of sewage, industrial wastes or other wastes into these waters. These waters shall contain no phosphorus and nitrogen in amounts that will result in growths of algae, weeds and slimes that will impair the waters for their best usages.
- Class A_{special}: Source of water supply for drinking, culinary or food processing purposes; primary and secondary contact recreation; and fishing. These waters shall be suitable for fish propagation and survival. These international boundary waters, if subjected to approved treatment equal to coagulation, sedimentation, filtration and disinfection, with additional treatment if necessary to remove naturally present impurities, will meet New York State Department of Health drinking water standards and will be considered safe and satisfactory for drinking water purposes
- Class AA: Source of water supply for drinking, culinary or food processing purposes; primary and secondary contact recreation; and fishing. These waters shall be suitable for fish propagation and survival. These waters, if subjected to approved disinfection treatment, with additional treatment if necessary to remove naturally present impurities, will meet New York State Department of Health drinking water standards and will be considered safe and satisfactory for drinking water purposes
- Class A: Source of water supply for drinking, culinary or food processing purposes; primary and secondary contact recreation; and fishing. These waters shall be suitable for fish propagation and survival. These waters, if subjected to approved treatment equal to coagulation, sedimentation, filtration and disinfection, with additional treatment if necessary to remove naturally

present impurities, will meet New York State Department of Health drinking water standards and will be considered safe and satisfactory for drinking water purposes

- Class B Suitable for primary and secondary contact recreation and fishing. These waters shall be suitable for fish propagation and survival
- Class C: Suitable for fishing, and fish propagation and survival. The water quality shall be suitable for primary and secondary contact recreation, although other factors may limit the use for these purposes.
- Class D: Suitable for fishing. Due to such natural conditions as intermittency of flow, water conditions not conducive to propagation of game fishery, or stream bed conditions, the waters will not support fish propagation. These waters shall be suitable for fish survival. The water quality shall be suitable for primary and secondary contact recreation, although other factors may limit the use for these purposes.
- Class (T): Designated for trout survival, defined by the Environmental Conservation Law Article 11 (NYS, 1984b) as brook trout, brown trout, red throat trout, rainbow trout, and splake

**APPENDIX C:
SUMMARY OF STATISTICAL METHODS USED TO EVALUATE TRENDS**

1. Non-Parametric Analyses

Kendall tau ranking orders paired observations by one of the variables (say arranging water clarity readings by date). Starting with the left-hand (say earliest date) pair, the number of times that the variable not ordered (in this case clarity readings) is exceeded by the same variable in subsequent pairs is computed as P, and the number of times in which the unordered variable is not exceeded is computed as Q. This computation is completed for each ordered pair, with N= total number of pairs (samples), and the sum of the differences $S = \Sigma(P-Q)$. The Kendall tau rank correlation coefficient **t** is computed as:

$$t = 2S/(N*(N-1))$$

Values for t range from -1 (complete negative correlation) to +1 (complete positive correlation). As above, strong correlations (or simply “significance”) may be associated with values for t greater than 0.5 (or less than -0.5), and moderate correlations may be associated with values for t between 0.3 and 0.5 (or between -0.3 and -0.5), but the “significance” of this correlation must be further computed. Standard charts for computing the probabilities for testing the significance of S are provided in most statistics text books, and for values of N greater than 10, a standard normal deviate D can be computed by calculating the quotient

$$D = S\sqrt{18} / \sqrt{[N(N-1)(2N+5)]}$$

and attributing the following significance:

$$D > 3.29 = 0.05\% \text{ significance}$$

$$2.58 < D < 3.29 = 0.5\% \text{ significance}$$

$$1.96 < D < 2.58 = 2.5\% \text{ significance}$$

$$D < 1.96 = > 2.5\% \text{ significance}$$

For the purpose of this exercise, 2.5% significance or less is necessary to assign validity (or, using the vernacular above, “significance”) to the trend determined by the Kendall tau correlation. It should be noted again that this evaluation does not determine the magnitude of the trend, but only if a trend is likely to occur.

Parametric trends can be defined by standard best-fit linear regression lines, with the significance of these data customarily defined by the magnitude of the best fit regression coefficient $\text{\textcircled{R}}$ or R^2). This can be conducted using raw or individual data points, or seasonal summaries (using some indicator of central tendency, such as mean or median). Since the former can be adversely influenced by seasonal variability and/or imprecision in the length and breadth of the sampling season during any given year, seasonal summaries may provide more realistic measures for long-term trend analyses. However, since the summaries may not adequately reflect variability within any given sampling season, it may be appropriate to compare deviations from seasonal means or medians with the “modeled” change in the mean/median resulting from the regression analyses.

When similar parametric and non-parametric tools are utilized to evaluate long-term trends in NYS lakes, a few assumptions must be adopted:

- Using the non-parametric tools, trend “significance” (defined as no more than appx. 3% “likelihood” that a trend is calculated when none exists) can only be achieved with at least four years of averaged water quality data. When looking at all summer data points (as opposed to data averaging), a minimum of forty data points is required to achieve some confidence in data significance. This corresponds to at least five years of CSLAP data. The “lesson” in these assumptions is that data trends assigned to data sets collected over fewer than five years assume only marginal significance.

As noted above, summer data only are utilized (as in the previous analyses) to minimize seasonal effects and different sampling schedules around the fringes (primarily May and September) of the sampling season. This reduces the number of data points used to compile averages or whole data sets, but is considered necessary to best evaluate the CSLAP datasets.

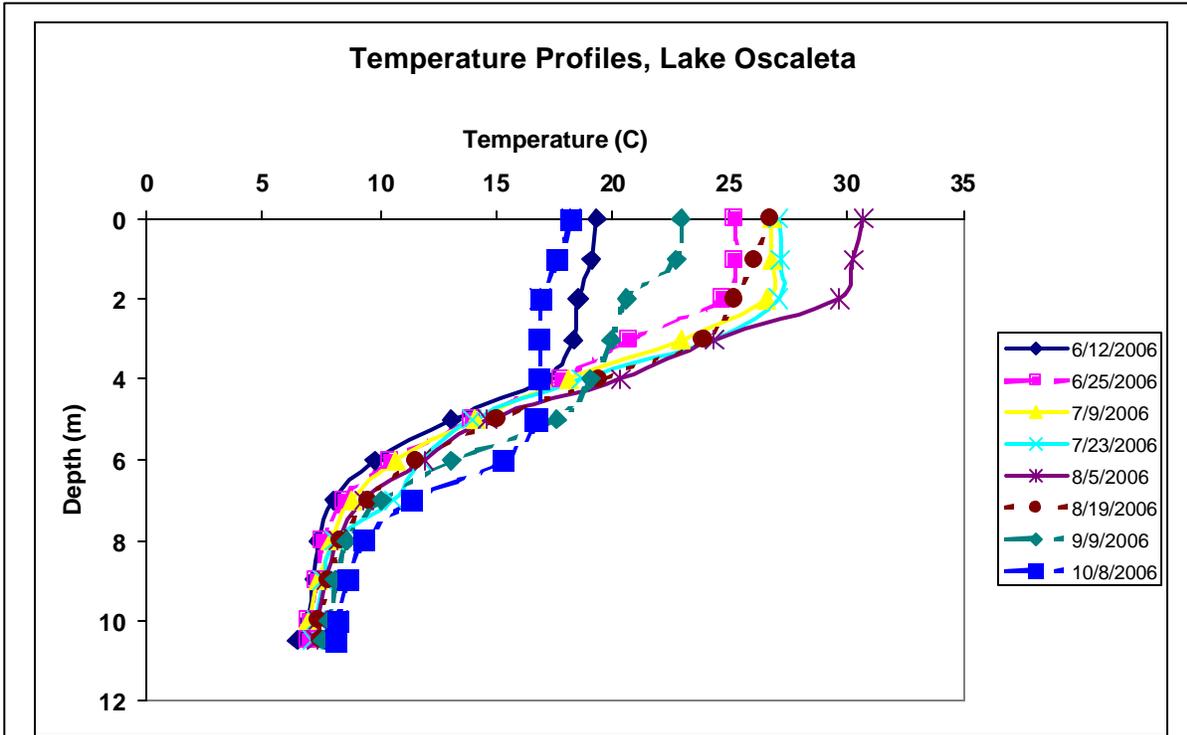
2. Parametric Analyses

Parametric analyses are conducted by comparing annual changes in summer mean values for each of the analyzed sampling parameters. Summer is defined as the period from June 15 thru September 15, and roughly corresponds to the window between the end of spring runoff (after ice out) and start of thermal stratification, and the onset of thermal destratification. This period also corresponds to the peak summer recreational season and (for most lakes) the most critical period for water quality impacts. It also bounds the most frequent range of sampling dates for the majority of both the primarily seasonal volunteers and full time residents of CSLAP lakes.

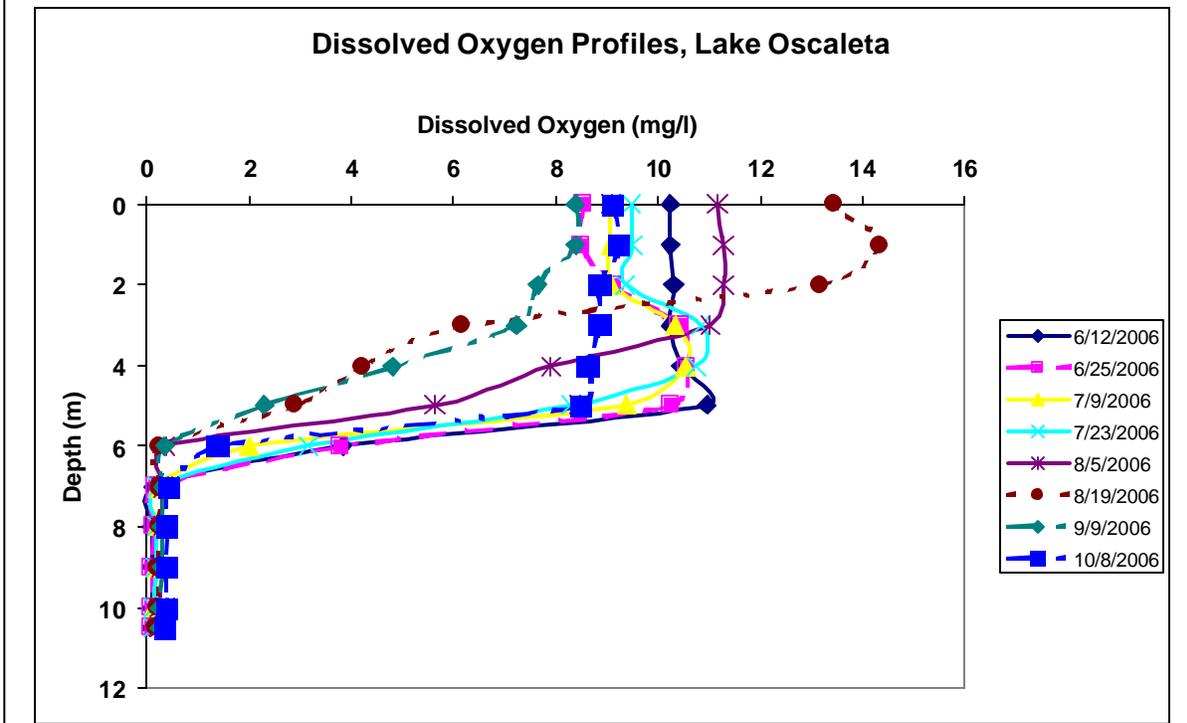
Trends in the parametric analyses are determined by the least squares method, in which “significance” requires both a high correlation coefficient ($R^2 > 0.5$) and intra-seasonal variance to be lower than the predicted change (trend) over the period of sampling (roughly corresponding to Δy). Changes in water quality indicators are also evaluated by the two-sided t-test, in which the change (z statistic) in the mean summer value for each of the indicators by decade of sampling (1980s, 1990s, 2000s) is compared to the t statistic distribution within the 95% confidence interval, with the null hypothesis corresponding to no significant change.

APPENDIX D: BACKGROUND INFO FOR LAKE OSCALETA

CSLAP Number	205
Lake Name	L Oscaleta
First CSLAP Year	2006
Sampled in 2005?	no
Latitude	411750
Longitude	733410
Elevation (m)	144
Area (ha)	23.3
Volume Code	13
Volume Code Name	Lower Hudson River
Pond Number	118
Qualifier	none
Water Quality Classification	B
County	Westchester
Town	South Salem
Watershed Area (ha)	4486000
Retention Time (years)	0.714285714
Mean Depth (m)	6.8
Runoff (m/yr)	0.56
Watershed Number	13
Watershed Name	Lower Hudson River
NOAA Section	5
Closest NOAA Station	Yorktown Heights
Closest USGS Gaging Station-Number	1374821
Closest USGS Gaging Station-Name	Titicus River at Purdys Station
CSLAP Lakes in Watershed	Burden L, Copake L, Duane L, Forest L-R, Hillside L, Indian L-P, Kinderhook L, L Carmel, L Celeste, L Lincolndale, L Lucille, L Mahopac, L Meahagh, L Mohegan, L Myosotis, L Nimham, L Oscaleta, L Oscawana, L Ossi, L Peekskill, L Rippowam, L Taghanic, L Tibet, L Truesdale, L Waccabuc, Long P, Monhagen L, Nassau L, Orange L, Peach L, Queechy L, Robinson P, Round P, Sagamore L, Sepasco L, Shaver P, Shawangunk L, Shenerock L, Snyders L, Spring L, Teatown L, Thompsons L, Tomkins L, Whaley L



Temperature Profiles at Lake Oscaleta, 2006



Dissolved Oxygen Profiles at Lake Oscaleta, 2006